

1 **Title:** Towards reproducible metabarcoding data - lessons from an international cross-laboratory
2 experiment

3 **Running title:** Towards reproducible metabarcoding data

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34

35 **Abstract**

36 Advances in high-throughput sequencing (HTS) are revolutionizing monitoring in marine environments by
37 enabling rapid, accurate and holistic detection of species within complex biological samples. Research
38 institutions worldwide increasingly employ HTS methods for biodiversity assessments. However, variance
39 in laboratory procedures, analytical workflows and bioinformatic pipelines impede the transferability and
40 comparability of results across research groups. An international experiment was conducted to assess the
41 consistency of metabarcoding results derived from identical samples and primer sets using varying
42 laboratory procedures. Homogenized biofouling samples collected from four coastal locations (Australia,
43 Canada, New Zealand and the USA) were distributed to 12 independent laboratories. Participants were
44 asked to follow one of two HTS library preparation workflows. While DNA extraction, primers and
45 bioinformatic analyses were purposefully standardized to allow comparison, many other technical variables
46 were allowed to vary among laboratories (e.g., amplification protocols, type of instrument used, etc.).
47 Despite substantial variation observed in raw results, the primary signal in the data was consistent, with the
48 samples grouping strongly by geographic origin for all datasets. Simple post-hoc data clean-up by removing
49 low quality samples gave the best improvement in sample classification for nuclear 18S rRNA gene data,
50 with an overall 92.81% correct group attribution. For mitochondrial COI gene data, the best classification
51 result (95.58%) was achieved after correction for contamination errors. The identified critical
52 methodological factors that introduced the greatest variability (preservation buffer, sample defrosting,
53 template concentration, DNA polymerase, PCR enhancer) should be of great assistance in standardizing
54 future comparative biodiversity studies using metabarcoding.

55

56 **Keywords:** reproducibility, high-throughput sequencing, 18S ribosomal rRNA (18S rRNA),
57 mitochondrial Cytochrome c Oxidase subunit 1 (COI), metabarcoding, standardization.

58

59 **Introduction**

60 Recent advances in high-throughput sequencing (HTS) are revolutionizing ecology, offering
61 unprecedented opportunities for new species discovery, monitoring ecological trends, diet analysis and
62 environmental impact assessment (Aylagas, Borja, & Rodrigues-Ezpeleta, 2014; L. J. Clarke, Trebilco,
63 Walters, Polanowski, & Deagle, 2020; Dowle, Pochon, Banks, Shearer, & Wood, 2016; Keeley, Wood, &
64 Pochon, 2018; Valentini et al., 2016; Zhan et al., 2013). Metabarcoding enables community-wide
65 biodiversity to be characterized from environmental DNA (eDNA) samples. It is based on the mass-
66 sequencing of short, conserved DNA fragments (molecular markers or barcodes), which are then assigned
67 taxonomy based on reference sequence databases, allowing the simultaneous characterization of a large
68 range of taxa (Hajibabaei, Shokralla, Zhou, Singer, & Baird, 2011; Taberlet, Coissac, Hajibabaei, &
69 Rieseberg, 2012). Despite known limitations (e.g., incomplete reference sequence databases, uncertainties
70 around sensitivity, specificity and detection probabilities), this approach is increasingly being advocated
71 as a rapid, cost-effective and sensitive tool for environmental monitoring and assessment (Cordier et al.,
72 2020; Lacoursière-Roussel et al., 2018; Wood et al., 2013). Metabarcoding can be used to detect all life-
73 stages, including morphologically indistinguishable larval forms and rare or sparsely distributed
74 populations, often with less time-effort than conventional approaches (Brown, Chain, Zhan, MacIsaac, &
75 Cristescu, 2016; Comtet, Sandionigi, Viard, & Casiraghi, 2015; Zaiko, Samuiloviene, Ardura, & Garcia-
76 Vazquez, 2015). In some situations, it offers greater potential for standardization than traditional
77 morphological biodiversity assessment (Aylagas, Borja, Irigoien, & Rodríguez-Ezpeleta, 2016; Porter &
78 Hajibabaei, 2018). However, metabarcoding is inherently complex, with many technical and analytical
79 steps that inevitably vary among users, as well as multiple reputable but distinct metabarcoding
80 workflows (Cristescu, 2014; Goldberg et al., 2016; Murray, Coghlan, & Bunce, 2015; Zaiko, Pochon,
81 Garcia-Vazquez, Olenin, & Wood, 2018). Virtually all steps in the workflow vary among laboratories and
82 individual studies, including sampling, preservation methods, DNA extraction, PCR amplification
83 protocols, library preparation methods, sequencing platforms and bioinformatic pipelines (Bailet et al.,

84 2020; Bowers et al., 2021). This creates a significant challenge for the standardization of protocols and
85 the implementation of a unified, cross-calibrated approach.

86 For sensitive applications such as biosecurity surveillance, a field devoted to monitoring
87 environments for species with potentially harmful ecological and economic consequences (Molnar,
88 Gamboa, Revenga, & Spalding, 2008; Simberloff et al., 2013), standards of quality assurance and control
89 are of higher importance than in general biodiversity studies (Darling & Mahon, 2011; Lehtiniemi et al.,
90 2015). Reporting unverified biosecurity risks from HTS data may lead to inadequate management responses
91 and even potential legal entanglements for researchers (Darling, Pochon, Abbott, Inglis, & Zaiko, 2020).
92 Despite recent calls for improved consistency in HTS-based research outputs (Jeunen et al., 2019; van der
93 Loos & Nijland, 2020), laboratories working in this field often follow different analytical workflows,
94 develop and apply in-house reference sequence databases, use customized bioinformatic pipelines and often
95 fail to report methodologically relevant metadata consistently (Goldberg et al., 2016; Nicholson et al.,
96 2020). This lack of harmonization may impede the reproducibility of research, transferability and
97 comparability of results, and ultimately, the credibility of HTS-based biodiversity assessments (Zaiko et
98 al., 2018). There are an increasing number of empirical studies addressing the effect of single components
99 of the workflow, such as filtration, eDNA extraction methods, PCR thermocycling conditions, choice of
100 polymerase or sequencing platform used (Aylagas et al., 2016; Braukmann et al., 2019; Djurhuus et al.,
101 2017; Jeunen et al., 2019; Nichols et al., 2018). However, what is still largely unknown is the degree to
102 which variations in common metabarcoding protocols distort community composition data and the impact
103 of these potential biases on interpreting of metabarcoding-derived biodiversity information.

104 Acknowledging the need for a coordinated effort to accelerate the uptake of HTS-based tools for
105 biodiversity surveys and conservation applications, an international cross-laboratory experiment was
106 conducted to evaluate replicability of a metabarcoding protocol and explore how laboratory-based variance
107 in sample handling and processing impacts biodiversity assessments. The overarching goal of the present
108 study was to investigate the consistency and reproducibility of metabarcoding results when a set of identical

109 environmental samples are analyzed simultaneously by different laboratories (12 in this study) following a
110 semi-standardized protocol and analytical workflow. The applied experimental design allowed us to
111 identify discrepancies caused by laboratory-specific variation in technical steps and find the weakest links
112 in the analytical pipelines that resulted in the greatest divergence in metabarcoding data. In our study we
113 hypothesized that: (1) despite laboratory-to-laboratory variations (e.g., technical equipment, PCR
114 conditions and reagents), consistent patterns of variability in biological community information would be
115 derived for two markers (COI and 18S) when using a semi-standardized protocol; (2) the use of different
116 HTS library build protocols (dual index primers *versus* fusion primers) would have no effect on the
117 community composition data but might impact levels of contamination; (3) the greatest variability would
118 be introduced in the PCR and clean-up steps, given that different reagents and equipment were used in each
119 laboratory; and (4) if variability in community composition was observed between laboratories, this could
120 be mitigated by applying post-hoc data correction.

121

122 **Methods**

123 An overview of the experimental set-up is presented in Figure 1, with additional details provided
124 in the sections below.

125 *Collection and processing of biofouling samples*

126 Mature marine biofouling (~20 mL) was collected from upper 0.5-1 m depth at international
127 locations adjacent to busy marine ports and marinas (Table 1). The samples were preserved with
128 RNAlater™ (ThermoFisher Scientific, USA) or LifeGuard™ (Qiagen, Venlo, The Netherlands) by
129 adding a volume of preservative solution equivalent to at least 150% of the biological sample volume
130 (Table 1).

131 All samples were delivered to the Cawthron Institute, New Zealand (transported in ambient
132 temperature within 2 days) and processed uniformly to ensure the same starting material was used by all

133 laboratories. Samples were homogenized by bead beating (1,500 rpm, 5 min; 1600 MiniG Spex
134 SamplePrep NJ, USA) and split into 4×0.5 mL aliquots per laboratory (192 aliquots in total, Figure 1).
135 Twelve well-established molecular laboratories in six countries participated in this study: New Zealand
136 (Cawthron Institute, National Institute of Water and Atmospheric Research Ltd, University of Auckland
137 and University of Otago); Australia (Curtin University and Macquarie University); USA (Moss Landing
138 Marine Laboratory); Canada (Department of Fisheries and Oceans, McGill University and University of
139 Guelph); Spain (University of Oviedo) and Lithuania (Klaipeda University).

140 *Environmental DNA isolation and sequencing library preparation*

141 Variation at the DNA extraction step was minimized by providing participants from each of the
142 12 laboratories with a MoBio/Qiagen PowerBiofilm DNA isolation kit (Qiagen, Venlo, The Netherlands)
143 and a detailed manufacturer's protocol for DNA isolation, that all participants were asked to follow.

144 Participating laboratories were randomly allocated to a treatment group that followed one of two HTS
145 library preparation workflows (dual index primers or fusion primers, see below) for amplification of both
146 the 18S rRNA (Zhan et al. 2013; Uni18S: 5'-AGGGCAAKYCTGGTGCCAGC-3', Uni18SR: 5'-
147 GRCGGTATCTRATCGYCTT-3') and COI (Leray et al. 2013; Geller et al. 2013; mlCOIintF: 5'-
148 GGWACWGGWTGAACWGTWTAYCCYCC-3', jgHCO2198: 5'-
149 TAIACYTCIGGRTGICCRAARAAYCA-3') genes. For analytical objectivity, the laboratories within
150 each group were randomly assigned a letter code – A to F for the dual index and G to L fusion groups,
151 respectively. All results are hereafter reported under those codes to avoid attributing analytical outcomes
152 to a particular laboratory.

153 Protocols for PCRs and library preparation were also provided (Supporting Information,
154 Appendices I and II); however, given that each laboratory uses different equipment and/or used their
155 preferred reagents, some flexibility in equipment and reagents was allowed at the PCR and clean-up steps.

156 The participants recorded any deviations from the provided protocol, as well as specific details on sample
157 handling and any issues encountered during the workflow (Table S1).

158 *High-throughput sequencing*

159 Libraries were sequenced on MiSeq™ Illumina instruments at two sequencing facilities:
160 University of Guelph, Canada (libraries from laboratories A-F, dual index primers, n = 120) and Curtin
161 University, Australia (libraries from laboratories G-L, fusion primers, n = 120). For the dual index library,
162 18S rRNA gene and COI amplicons were pooled for each sample and indexed using Illumina's Nextera
163 XT v2 index adapters (FC-131-2001). Indexed samples were pooled and purified using AMPure XP
164 magnetic bead kit (Beckman Coulter). The library concentration was quantified with Invitrogen
165 PicoGreen dsDNA assay kit on a TBS-380 fluorometer (Turner Biosciences). Fragment length was
166 assessed on an Agilent Bioanalyzer using Agilent's DNA 7500 kit. The library was diluted to 4nM and
167 sequenced on an Illumina MiSeq using a v3 chemistry kit (2 × 300). Each run had a 10% PhiX spike-in.
168 Samples were sequenced over two runs: Run 1 contained groups A, D and F. Run 2 contained groups B,
169 C, and E.

170 For the fusion library, PCR products were quantified on arrival at the sequencing facility with
171 Qubit (ThermoFisher Scientific, USA) and blended at equal molarity for three sequencing runs: Run 1 -
172 18S rRNA gene and COI from laboratories G and L; Run 2 - 18S rRNA gene and COI from laboratories
173 H and K; Run 3 - 18S rRNA gene and COI from laboratories I and J. Amplicons from the laboratory J
174 contained traces of green dye from the Taq buffer and were further cleaned using a QiaQuick Purification
175 Kit (QIAGEN, Carlsbad, USA), and eluted in 30 µL elution buffer before quantification and library
176 pooling. Final libraries were quantified again using Qubit. Molarity was calculated and diluted to 2 nM,
177 except for Run 1, which had concentration of 1.3 nM. Libraries were denatured and blended with 20 pM
178 PhiX. Sequencing was performed on the MiSeq (Illumina) system with a v2 chemistry kit (2 × 250).

179 *Bioinformatics*

180 Bioinformatics analysis of all the 18S rRNA and COI datasets was conducted at the
181 Commonwealth Scientific and Industrial Research Organization (CSIRO, Australia), using the GHAP
182 amplicon pipeline (<https://doi.org/10.4225/08/59f98560eba25>). This is a conventional amplicon clustering
183 and classification pipeline built around tools from USearch (Edgar, Haas, Clemente, Quince, & Knight,
184 2011), combined with in-house scripted tools for demultiplexing and generating OTU tables. The
185 amplicon reads were demultiplexed, split and trimmed as needed and the read pairs were then merged and
186 de-replicated. The merged reads were trimmed/size-selected and UNOISE3 (Edgar, 2016) used to group
187 almost identical sequences to generate zOTUs (Zero-radius Operational Taxonomic Units). Each zOTU
188 sequence was then classified by BLASTing against a curated set of reference sequences. The 18S rRNA
189 gene zOTUs were matched against curated sequences derived from the SILVA v128 SSU reference set
190 (Quast et al., 2013). The COI zOTUs were matched against a set of COI reference sequences downloaded
191 from GenBank (on 24-9-2019). The pipeline then mapped the merged reads back onto the zOTU
192 sequences to get accurate read counts for each zOTU/sample pair, and generated zOTU abundance tables
193 in both text and .biom (v1) formats, complete with taxonomic classifications and species assignments.
194 The zOTU abundance tables were then summarized at all resolved taxonomic levels, combining the
195 counts for identified taxa across all OTUs.

196 Both the dual index and fusion library sequencing was undertaken on pooled samples, so each
197 data file contained both 18S rRNA gene and COI data. These files were split into 18S rRNA gene and
198 COI files based on the primer sequences found at the start of each of the reads, and these primer regions
199 were then trimmed from the reads. After splitting, there were found to be unequal numbers of 18S rRNA
200 gene and COI reads in both datasets, with only 6-15% (fusion) and 30% (dual index) of the samples being
201 assigned to the 18S rRNA gene.

202 The fusion sequencing protocol also added 6 or 7 bp barcode to the start of each read, with these
203 barcodes removed from the reads during demultiplexing. The MiSeq runs used for the fusion libraries
204 produced 260 bp reads (including barcodes and primers), and demultiplexing/trimming these resulted in

205 reads of 226-228 bp (18S rRNA) and 232-234 bp (COI). The dual index MiSeq runs produced 301 bp
206 reads, including primer sequences, and used standard Illumina out-of-read barcodes. The number of bases
207 removed by this post-sequencing trimming step varied from sample to sample, with higher levels of
208 trimming observed for 18S rRNA reads (median 10.5 bp) than for COI (median 0.7 bp) reads. Samples
209 from Lab C were more heavily trimmed than other samples (average trim for 18S rRNA was 34.5 bp, and
210 75.3 bp for COI). After splitting and primer-trimming, the reads going into the amplicon pipeline were
211 15-282 bp long (mean 246-280 bp) for 18S rRNA gene, and 9-275 bp (mean 200-275 bp) for COI. The
212 inter-primer region being sequenced was ~410 bp for 18S rRNA gene, and ~310 bp for COI.
213 Consequently, both the dual index and fusion library reads were long enough to have sufficient overlap
214 for satisfactory pair-merging.

215 *Statistical analyses*

216 For each zOTU table (18S rRNA gene and COI) resulting from the bioinformatics pipeline, a
217 permutational multivariate analysis of variance (PERMANOVA, Anderson 2001, 2017) was used to test
218 how variation in HTS-derived biodiversity was partitioned in response to ‘sample origin’, ‘library’ (dual
219 index vs fusion; both fixed factors) and ‘laboratory’ (a random factor nested in ‘library’). Analysis was
220 performed on Bray-Curtis similarity matrices of fourth root transformed datasets (zOTU reads data),
221 applying 999 permutations of residuals under a reduced model with type III sum of squares applied. To
222 assess the relative spread of data clouds in samples grouped by origin, a test for homogeneity of
223 multivariate dispersions (PERMDISP) was applied for each dataset with 999 permutations. Individual
224 sample distances to origin group centroids were retrieved for library type, 18S rRNA gene and COI
225 separately. A distribution-free Wilcoxon signed-rank test was used to check whether 18S rRNA and COI
226 OTU richness differ significantly between libraries, as well as whether distances to group centroids differ
227 significantly between 18S rRNA and COI datasets.

228 To infer which factors affected the consistency of biodiversity information during sample
229 handling and processing, a multiple regression model with automated stepwise selection function was

230 implemented in R Stats package version 3.5.3 (R-project, 2014). In the model, distance of individual
231 samples to centroids of the sample origin group was treated as response variable, and the following
232 sample handling and processing parameters were considered as predictor variables: preservation buffer
233 (RNAlater/LifeGuard), storage time before DNA extraction (months), multiple (more than two times)
234 DNA defrosting cycles due to PCR troubleshooting (yes/no), library type (fusion/dual index), amount of
235 template used in PCR (μL), dilution of DNA template (yes/no), DNA polymerase used, addition of
236 Bovine Serum Albumin (BSA) or other enhancers (yes/no), number of PCR cycles, contamination
237 detected in control samples (yes/no). For each gene and library type we also performed correlation
238 analysis using cor.test function implemented in R to test for association (Pearson product-moment
239 correlation coefficient) between sequence Phred quality score (averaged per sample) and distance to
240 origin group centroids.

241 For each dataset (18S rRNA gene and COI), a series of Canonical Analyses of Principal
242 Coordinates (CAP) with 'sample geographic origin' as a grouping factor and 999 permutations were
243 applied to assess the classification success (correct allocation of samples into a pre-defined group) on raw
244 (uncorrected) HTS data and following post-hoc corrections: **Corr1**) samples with low sequence number
245 ($<1,000$) excluded; **Corr2**) samples randomly rarefied to 10,000 sequence depth; **Corr3**) OTUs found in
246 the negative controls removed from the corresponding subsets of data; **Corr4**) maximum read abundance
247 of OTUs found in the negative controls subtracted from the corresponding subsets of data; **Corr5**) OTUs
248 found in negative controls subtracted and samples with low sequence number ($<1,000$) excluded; **Corr6**)
249 OTUs found in negative controls were subtracted and samples rarefied to 10,000 sequence depth.
250 Clustering of samples according to their origin was visualized by CAP ordination plots for the raw dataset
251 and corrected dataset with the best classification success. All multivariate statistical analyses were carried
252 out using PRIMER 7 (v 7.0.13) with the PERMANOVA + add-on (K. R. Clarke & Gorley, 2015).
253 Following dataset corrections which resulted in the best 18S rRNA and COI sample classification
254 according to the origin, the laboratory and library-driven variability in biodiversity data were explored

255 using Principal Coordinates Analysis (PCoA). For each sample origin, 18S rRNA and COI datasets, a
256 two-dimensional visualization was performed using plot_ordination function on Bray-Curtis similarity
257 matrices in phyloseq R package (McMurdie & Holmes, 2013).

258 **Results**

259 *General observations*

260 Although all 12 laboratories strictly adhered to the provided DNA extraction protocol, large
261 variations in DNA concentration and quality were observed (Figure S1). However, different instruments
262 were used to estimate DNA concentrations (Table S1) and some participants expressed concerns about
263 accuracy of their measurements. The quantity and quality of extracted DNA is not a direct indication of
264 how the targeted eukaryotic genes will amplify, as bulk samples are likely to comprise large proportions
265 of bacterial DNA (not targeted in this study). Therefore, DNA concentration data were not used in
266 subsequent statistical analyses.

267 At the PCR and clean-up steps, all laboratories applied variations to the suggested protocols.
268 Seven laboratories reported difficulties in amplifying one or both marker genes and had to troubleshoot
269 by adjusting cycling conditions, trying different reagents and/or template concentrations (Table S1). A
270 variety of positive controls were used in library preparation by different laboratories, including DNA
271 from a single species, or mock communities derived from combined DNA of multiple species (see Table
272 S1).

273 The HTS from 240 amplicons (120 from the dual index and 120 from the fusion libraries)
274 resulted in 15,104,359 and 41,241,053 pair-end sequence reads for dual index and fusion primers library,
275 respectively. The total number of quality-filtered, de-noised (non-chimeric) sequences derived from 18S
276 rRNA gene amplicons were 6,231,102 and 4,211,218 from dual index and fusion primers library, and
277 7,776,009 and 27,540,509 for COI amplicons from dual index and fusion primers library, respectively.
278 Two samples from the 18S rRNA gene dataset failed to produce any high-quality sequence reads and

279 were therefore excluded from downstream analyses. The average sequence read Phred quality was
280 relatively consistent for 18S rRNA and COI gene amplicons, for both library types (Figure S2). However,
281 the sequence quality of the COI fusion library samples was, overall, slightly higher for both forward and
282 reverse reads.

283 High variability in the number of filtered sequences and OTUs across samples and laboratories
284 was observed (Figures 2 and 3). In the fusion library group, the number of good quality sequences yielded
285 per sample varied from 1 to 50,101 for 18S rRNA gene amplicons and 347 to 556,775 for COI amplicons,
286 respectively. In the dual index library, differences in sequence yields between genes were similar, with
287 data from the 18S rRNA gene ranging from 1 to 189,779 and COI from 6 to 219,694. The number of
288 retrieved OTUs in the dual index library ranged from 1 to 465 and 4 to 1,622 for 18S rRNA gene and COI
289 markers, respectively, and showed significant ($p < 0.001$) moderate correlation with the number of
290 sequence reads ($r = 0.52$ and 0.68 , 18S rRNA and COI samples, respectively). In the fusion library, OTU
291 numbers correlated only with sequence reads for the 18S rRNA samples (0.46 , $p < 0.001$), ranging from 1
292 to 436 OTUs per sample. COI samples were characterized by an overall higher number of OTUs (79 to
293 2,363 per sample), and this did not correlate with the number of sequence reads ($r = 0.18$, $p = 0.08$). The 18S
294 rRNA OTU richness did not differ significantly between libraries ($W = 4975$, $p = 0.22$), however overall
295 higher COI OTU number was observed in the fusion library ($W = 2338$, $p < 0.001$). We observed no
296 difference between the two markers ($W = 19429$, $p\text{-value} = 0.2707$) in relative mean spread of data clouds
297 in samples grouped by origin between the two markers (average distance to group centroids were $53.89 \pm$
298 $SD 12.93$ and $52.98 \pm SD 15.14$ for 18S rRNA and COI respectively).

299

300 Three laboratories reported contamination in their negative control samples, while sequencing
301 data analysis revealed substantial contamination in DNA extraction blanks in six dual index datasets
302 (Figure 4). Species used as positive controls were successfully retrieved from 17 out of 24 datasets (i.e.
303 70.8%). Failed positive controls affected three dual index and four fusion samples (Figure 4). None or
304 negligible contamination was detected in the negative (no template) PCR controls (except for laboratory
305 D, where >10,000 sequences of the same OTU found in extraction blanks were detected).

306 *Biodiversity clustering according to sample origin – nuclear small subunit (18S rRNA gene) dataset*

307 All factors considered in the PERMANOVA design (sample origin, library type and laboratory),
308 as well as their interactions, had a significant effect on the partitioning variation in 18S rRNA gene
309 datasets ($p < 0.05$, Table S2). A test of homogeneity for the within-group multivariate dispersion
310 (PERMDISP) revealed significant differences in the average distances to centroids among samples
311 grouped by sample origin ($p = 0.001$, $F = 18.869$, PERMDISP).

312 The multiple linear regression followed by an automated stepwise selection of optimal model
313 showed that ‘preservation buffer’ ($p < 0.001$), type of DNA polymerase used in PCR reactions ($p < 0.001$),
314 addition of PCR enhancers ($p < 0.001$) and ‘multiple defrosting of DNA samples’ ($p < 0.001$) all had
315 significant effects on the dispersion around the origin group centroid (Table S3). The best selected model
316 explained 46% of variation in distance values. In dual index library samples, distances to group centroids
317 showed weak ($r = -0.35$ and -0.34) but statistically significant ($p < 0.001$) negative correlations with the
318 sequence quality scores (forward and reverse reads respectively, Figure S3). No such correlations were
319 detected in fusion library samples ($p > 0.5$; $r = 0.11$ and -0.12 , forward and reverse reads respectively,
320 Figure S3).

321 Canonical analysis of principal coordinates (CAP) models derived from either the raw 18S rRNA
322 dataset or datasets with post-hoc corrections were all characterized by high and significant canonical
323 correlations (Table 2, Figure 4), and correctly allocated 88% of samples by their origin in the uncorrected

324 data and up to 93% after applying post-hoc corrections. This is considerably better than the 25% success
325 expected by chance if samples were randomly allocated into four groups. Overall, the correct
326 classification of samples by their geographic origin was improved most by removing samples with
327 extremely low number of sequences (<1,000, Corr1, Table 2, Figure 4). Other corrections applied to the
328 dataset had variable success in improving classification. The AUS samples were classified most
329 accurately after rarefying sequencing data to equal depth across all samples (Corr2 and Corr6, Table 2).
330 The NZ samples showed the lowest classification success overall, not exceeding 87.5% correct allocation
331 (Corr1 and Corr6, Table 2).

332 *Biodiversity clustering according to the geographic origin - mitochondrial Cytochrome c Oxidase subunit*
333 *1 dataset*

334 All factors, including the interaction between origin and laboratory had significant effects
335 ($p < 0.01$, Table S4) on the partitioning variation in the COI datasets. Test of homogeneity for the within-
336 group multivariate dispersion (PERMDISP) revealed significant differences in distances to centroids
337 among the samples grouped by origin ($p = 0.01$, $F = 4.569$, PERMDISP).

338 Multiple linear regression followed by an automated stepwise selection of optimal models
339 showed a significant effect of the amount of DNA template and the type of DNA polymerase used in PCR
340 reactions ($p < 0.001$), 'preservation buffer' ($p < 0.001$), and addition of PCR enhancers ($p < 0.001$) on the
341 HTS data spread around group centroids (Table S5). The best selected model explained 58% of variation
342 in distance values. Statistically significant ($p < 0.001$) but moderate negative correlations between
343 distances to group centroids and sequence quality scores were detected in dual index samples ($r = -0.64$
344 and -0.68 , forward and reverse reads respectively, Figure S4). No such correlations were detected in
345 fusion library samples ($p > 0.5$; $r = 0.06$ and -0.16 , forward and reverse reads respectively, Figure S4).

346 Canonical analysis of principal coordinates (CAP) analyses of COI datasets yielded high and significant
347 canonical correlations (Table 3, Figure 5), and correctly allocated 88% of samples by their origin in the

348 raw datasets and up to 95% after applying post-hoc corrections (Table 3). The largest improvement in
349 classification was after corrections for contamination effects (Corr3, Corr4 and Corr5, Table 3). New
350 Zealand samples showed the lowest classification success, except for Corr3, when all OTUs found in
351 negative controls were removed.

352 *Variability in community biodiversity of the corrected datasets within the sample geographic origin*
353 *groups*

354 A more detailed exploration of biodiversity patterns following corrections that resulted in the
355 most improved sample classification by origin (Corr1 and Corr3 for 18S rRNA and COI datasets,
356 correspondingly) showed some variability within each location (Figures 6 and 7). It revealed variance
357 between dual index and fusion library datasets, which was more apparent in 18S rRNA data. However,
358 the first PCoA axis that separated the two library types explained only 19-24% and 10-24% of the total
359 variance, in 18S rRNA and COI datasets, respectively.

360 Most samples clustered reasonably well by laboratory, however this clustering differed across
361 datasets indicating stochasticity between sample origin groups. Some samples (e.g., fusion library
362 samples from laboratories H and I) formed rather consistent outlying groups, which affected dispersion
363 between fusion and dual index groups. This was particularly evident in the 18S rRNA dataset. Samples
364 from these laboratories were characterized by comparatively high average sequence read quality scores
365 (Lab H: $30.88 \pm \text{SD } 2.32$ and $34.84 \pm \text{SD } 2.40$; Lab I: $33.95 \pm \text{SD } 1.27$ and $35.07 \pm \text{SD } 1.79$, for 18S
366 rRNA and COI sequences respectively), good recovery of target taxa in the positive control samples
367 (except for 18S rRNA Lab H sample) and they did not encounter any contamination issues (Figure 4). In
368 terms of the sample processing workflows, the only distinct difference in Labs H and I from other
369 laboratories following fusion library protocol, was the addition of bovine serum albumin (BSA) at the
370 PCR step (see Table S1).

371

372 **Discussion**

373 The increased use of environmental DNA (eDNA) metabarcoding across freshwater, marine, and
374 terrestrial ecosystems is a result of advances in HTS technologies and their increasing affordability.

375 Almost anyone and anywhere, including citizen scientists, can now collect a sample for DNA extraction
376 and send it to a specialized laboratory to examine biodiversity in different biological matrices (Evans et
377 al., 2016; Jeunen et al., 2019; Larson et al., 2020; Vandamme et al., 2016). This offers unprecedented
378 opportunities for integrative analyses and comparisons of biodiversity patterns across wide temporal and
379 spatial scales. However, as our findings clearly illustrate, there is a critical need for consistency and
380 accuracy throughout the laboratory processing and data analysis workflows.

381 Metabarcoding (like any other biodiversity assessment technique) is prone to taxon identification
382 errors, biases and contaminations that can result in false negatives and positives, and these can occur at
383 any point in the sample processing and analytical steps (Doi et al., 2019; Ficetola et al., 2015). Improving
384 knowledge of potential confounding factors and sources of variation is critical for increasing the
385 credibility and reproducibility of metabarcoding approaches and for enabling large scale comparative
386 studies and long-term monitoring (Baker, 2016). In the absence of standardized sample and analysis
387 processing protocols for HTS metabarcoding, it is often assumed by default that reproducible data can be
388 generated by closely following a credible workflow if it has been published in a peer-reviewed article.
389 However, as demonstrated in the recent review by Nicholson et al. (2020), many studies do not report
390 sufficient details to allow replication and/or appropriate data comparison between studies. It is also not
391 usual practice to report troubleshooting incidents or small adjustments made to the protocols, which are
392 common and evident in the present study. The importance of such variation on the outcome, therefore,
393 remains largely unknown and is likely underestimated.

394 Our study revealed considerable variation in metabarcoding results from similar biodiversity
395 analyses undertaken in different laboratories. However, the crucial questions addressed here were whether
396 this methodological variation obscured the true biological signal, what factors introduced the greatest

397 variance, and whether the variance could be effectively reduced via appropriate data filtering. The raw
398 data from all laboratories showed strong and consistent differentiation of biological samples from
399 different geographical regions, for both the 18S rRNA and COI genes. We acknowledge, however, that
400 the consistency in community composition clustering was assessed here at a broad geographical scale,
401 and the effect of derived biases may be more significant when depicting more subtle (i.e. regional)
402 biodiversity patterns.

403

404 *Patterns in variability of biological community information derived from both markers and both library*
405 *types*

406 Our results showed no significant difference between 18S rRNA and COI data spread around
407 geographic origin groups. This suggests that neither marker is more susceptible to differences in
408 laboratory processing protocols and associated data biases. The 18S rRNA marker is more conserved than
409 the COI, which typically results in comparatively lower species resolution and underestimation of species
410 richness (Leray & Knowlton, 2016; Pearman et al., 2021; von Ammon et al., 2018). This may explain our
411 observation of slightly stronger clustering of 18S rRNA data points according to the sample origin and
412 substantial improvement of classification of 18S rRNA dataset following the removal of low-quality
413 samples (i.e., those with only a small number of resulting sequence reads). This was in contrast to the
414 COI data, which showed the best improvement in classification after removal of contaminating
415 sequences. The reduced phylogenetic resolution in 18S rRNA data potentially makes it more resilient to
416 contamination.

417 Despite a significant effect of library type on variance partitioning in samples detected by
418 PERMANOVA (contributed only 2 and 13% to variation in 18S rRNA and COI samples, respectively)
419 neither regression nor CAP analyses showed significant differentiation of samples between fusion and
420 dual index libraries. However, an interesting interaction between library type and marker gene was
421 observed in the numbers of sequences: the 18S rRNA yielded more sequence reads using dual index
422 compared with the fusion method, and there was an opposite trend in the COI data.

423 Unlike dual index libraries, where amplicons are tagged with short nucleotide sequences using
424 commercial kits at the additional PCR step (Bourlat, Haenel, Finnman, & Leray, 2016), in the fusion
425 primers approach a target DNA fragment is simultaneously amplified and tagged with a long nucleotide
426 tail (Zizka, Elbrecht, Macher, & Leese, 2019). The one-step tagging of amplicons from multiple samples
427 with fusion primers was developed to improve cost- and time-efficiency of sample processing for
428 metabarcoding (Elbrecht & Leese, 2015; Elbrecht & Steinke, 2018; Stat et al., 2017; Zizka et al., 2019).
429 However, it may potentially introduce substantial biases into inferred biodiversity due to tag-specific
430 mismatches with the PCR template and subsequent variation in primer-binding efficiency between
431 taxonomic groups (O'Donnell, Kelly, Lowell, & Port, 2016). The biases of fusion primers reported
432 however, refer to variation in sequence abundance across replicates and not in the detection of taxa, which
433 is comparable between methods (O'Donnell et al., 2016; Zizka et al., 2019). Interestingly, in our study we
434 found a higher richness of OTUs for the COI gene when using the fusion approach over the two-step
435 approach, which contrasts to what one would expect if PCR inefficiencies resulted from longer primers.
436 Regardless, the two library building methods produced similar community profiles. However, the fact that
437 no effect of the average Phred scores on data dispersal was observed for both markers in the fusion library
438 (Figs. S3 and S4), suggests that stochastic biases not directly related to the quality of sequences may be in
439 play when applying the one-step PCR approach. It is also presumed that the dual indexing protocol is less
440 susceptible to inhibitors, resulting in improved amplification of complex samples, compared to the one-
441 step fusion approach (Zizka et al., 2019). While we did not expect varying levels of inhibiting substances
442 between two markers (they were run on the same set of biofouling samples), longer (~410 bp) 18S rRNA
443 amplicons might have been more prone to inhibition (McCord, Pionzio, & Thompson, 2014; Opel,
444 Chung, & McCord, 2010). This likely explains overall lower number of sequences yielded for 18S rRNA
445 with a higher number of sequences in the dual index library, as well as the best data consistency
446 improvement following removal of 23 samples with extremely low sequence numbers (<1,000). The
447 differential inhibition effect might also explain the outliers in the corrected 18S rRNA fusion dataset.
448 These outliers came from samples where PCR was undertaken with addition of BSA, a protein known to

449 relieve amplification inhibition (Kreader, 1996). This observation suggests that the effect of BSA on
450 metabarcoding results and especially obtaining true diversity needs further exploration, preferably in
451 dedicated controlled experiments.

452 In contrast, the overall number of COI sequences was consistently higher than 18S rRNA
453 sequences under both dual index (~1.5 times higher on average) and, particularly, fusion (~25 times
454 higher on average) approaches. Because the 18S rRNA and COI gene fragments were pooled and
455 sequenced simultaneously, the most parsimonious explanation is that the shorter gene (~310 bp) COI
456 amplicons were subjected to preferential selection on the Illumina flow cell (Engelbrektson et al., 2010).
457 Future studies will be required to test this hypothesis further, as well as potential implications of gene
458 multiplexing on the derived biodiversity information.

459 Nevertheless, a clear advantage of the one-step fusion library approach (besides its cost/time-
460 efficiency) is that it is less prone to cross-contamination (Zizka et al., 2019). This was observed in our
461 study results, where substantial laboratory contamination was detected in the dual-index library samples,
462 with a significant effect on data clustering. This was especially evident in the COI dataset, where the best
463 classification results were achieved following corrections for contamination effects. The lowest
464 classification success was observed in the NZ COI dataset, and was also likely due to the dominance of
465 contaminating OTUs in the NZ sample from laboratory F (76% of all sequences were removed from their
466 dataset at the post-hoc correction step).

467 In this study we considered the two sequence library preparation methods most employed by
468 participating laboratories. Another common workflow not included here is the ‘tagged PCR’ approach,
469 which includes PCR amplifications with relatively short metabarcoding primers carrying 5’ nucleotide
470 tags and subsequent ligation of adaptors (Binladen et al., 2007; Carøe & Bohmann, 2020). This method
471 has been recently touted as an effective alternative for metabarcoding library preparation, especially when
472 tag-jump free protocols are implemented to prevent false sequence assignment to samples and mitigate
473 chimera formation (Carøe & Bohmann, 2020). Additional comparative analysis is required to investigate
474 the robustness of this approach in the context of laboratory-induced variations.

475

476 *The effect of sample handling and processing factors on variability in metabarcoding data*

477 Analysis of the relative importance of considered random effects in introducing variability into
478 the metabarcoding results showed that both 18S rRNA and COI datasets were most affected by the type
479 of preservation buffer, type of DNA polymerase and the addition of PCR enhancers. Previous studies
480 have shown that the choice of sample preservation method can have a substantial effect on the integrity of
481 eDNA and may alter HTS-derived community structure to a greater or lesser degree (Gray, Pratte, &
482 Kellogg, 2013; Lee, Adams, & Klassen, 2019; Tatangelo, Franzetti, Gandolfi, Bestetti, & Ambrosini,
483 2014). Therefore, although not always practical, immediate processing or cold storage (ideally at -20°C or
484 lower temperatures) remains the gold standard for retaining utmost quality of genetic material in samples
485 and reducing risks of preservation-related biases in biodiversity assessment (Lee et al., 2019; Renshaw,
486 Olds, Jerde, McVeigh, & Lodge, 2015). However, on-site freezing is not always possible, and the use of
487 DNA/RNA isolation buffers remains a popular alternative for samples collected in remote locations or
488 when samples are to be shipped internationally. In the present study, we chose two commonly used
489 commercial solutions, the RNAlater™ and LifeGuard™ isolation buffers for stabilizing the biofouling
490 samples. The main criteria for this choice were i) the compatibility with the DNA extraction protocol, ii)
491 the effective DNA stabilization capacity over a range of temperatures, and iii) no restrictive chemicals in
492 relation to shipping. Since a number of studies have previously shown no substantial differences in DNA
493 preservation efficacy between the two buffers (Gomez-Silvan et al., 2018) and they are referred to as
494 equivalent in standardized genetic and genomic protocols (Duran & Cravo-Laureau, 2017; Hampton-
495 Marcell, Frazier, Moormann, Owens, & Gilbert, 2017), we allowed sample providers to select either. The
496 somewhat higher variance in metabarcoding data from the LifeGuard™ samples was largely determined
497 by the presence of a few outliers. Taking into account the complexity of the sampled matrices (von
498 Ammon et al., 2018), the observed effect of preservation solution could be related to sample-specific
499 biases driven by the presence and concentration of inhibiting substances, or the prevalence of hard-shelled
500 organisms reducing homogenization of the material. Therefore, we cannot confidently infer differential

501 performance of the two buffers for sample preservation based on the current results. These results
502 highlight the need to ensure that variables such as preservation buffer type, temperature and storage time
503 should be recorded and reported in the metabarcoding protocols and considered when analyzing data from
504 multiple sample sets, from different laboratories or collected over different timeframes.

505 DNA polymerases with proofreading activity (e.g., MyFi™ DNA Polymerase) may enhance the
506 formation of chimeric sequences compared to Taq DNA polymerases (e.g., AmpliTaq Gold™ DNA
507 Polymerase) due to the earlier occurrence of PCR saturation (Ahn, Kim, Song, & Weon, 2012; Judo,
508 Wedel, & Wilson, 1998). Among the six distinct DNA polymerases used in the present study, MyFi was
509 the only proofreading polymerase and it was used by five of the participating labs (C, G, H, K, and L).
510 We did not find any clear evidence showing that the observed variability related to ‘polymerase’ effect
511 was specifically linked to the use of a proofreading polymerase. Therefore, a more parsimonious
512 explanation is that the random effect was driven by somewhat differential performance of the five Taq
513 DNA polymerases used.

514 The PCR enhancer Bovine Serum Albumin (BSA) was added by four laboratories (A, E, H and I)
515 and appeared to have an impact on the derived metabarcoding outputs. As noted above, this is possibly
516 due to its ability to bind to inhibitory substances during PCR reactions, preventing inhibitory interactions
517 with DNA polymerase (Woide, Zink, & Thalhammer, 2010). Biofilms and biofouling material contain
518 many known PCR inhibitors, such as glycogen, polysaccharides and slat (Schrader, Schielke, Ellerbroek,
519 & Johne, 2012). Therefore, a certain level of inhibition was reasonably expected in our samples. The
520 results from the analyses suggest that the addition of BSA might have substantial effect on both variance
521 and composition of the detected communities (as evidenced by the outliers from corrected datasets of
522 Labs H and I). Because this study was not designed to fully investigate the impact of BSA and only a few
523 laboratories included enhancers, further in-depth studies are required to understand better the patterns of
524 BSA effects on derived biodiversity.

525 The 18S rRNA data also showed sensitivity to multiple freeze-thaw cycles of the samples. The
526 available (mostly anecdotal) evidence of the effect of multiple defrosting on DNA integrity, suggests that

527 even one cycle of freeze-thawing can reduce eDNA signal (Bowers et al., 2021). However, it is rarely the
528 case that metabarcoding library construction is achieved without repetition of the PCR step (and thus
529 multiple freeze-thawing of DNA material) for at least some samples (e.g., troubleshooting and adjusting
530 amplification conditions for problematic samples). This parameter is difficult to control for, and it is
531 rarely (if at all) reported in metabarcoding protocols. It would be practical, however, to keep track of and
532 report freeze-thaw cycles of DNA samples, reduce their number where possible and, most importantly,
533 divide DNA into multiple aliquots upon extraction.

534 The dispersion of the COI data from different laboratories was also affected by the amount of
535 template DNA used in PCR (unlike 18S rRNA data). Low quantities of template DNA and stochasticity
536 in early PCR cycles is known to affect reproducibility of metabarcoding results (Alberdi, Aizpurua,
537 Gilbert, & Bohmann, 2018; Leray & Knowlton, 2017). This effect is particularly inherent for primers
538 with reduced specificity (e.g., COI primers comprising a few degenerate bases to allow better matching
539 the mutationally saturated target regions) applied to highly diverse environmental samples (e.g., marine
540 samples), and can be reduced by performing multiple PCR technical replicates (Collins et al., 2019;
541 Ficetola et al., 2015). Technical replication was advised in our suggested sample processing protocols,
542 however, it was only followed by five laboratories: A, C (dual-index group) and G, H and L (fusion
543 group). There is still lack of agreement in the literature whether technical replication is necessary for
544 improved capture of “true” biodiversity (Marotz et al., 2019), especially considering the associated
545 substantial increase in time and cost effort. Technical replication might be advised for studies where true
546 (environmental) replication is restricted for some reasons, to mitigate the effect of within sample
547 variability.

548

549 *Effect of the post-hoc data correction on the observed variability in community composition*

550 The datasets we dealt with in our analyses were somewhat atypical. We purposefully included all
551 investigated sample sets, including those that amplified poorly, had low numbers of sequence reads, and

552 controls with some contamination, which would have usually been excluded in a typical study. These
553 “problematic” samples most likely would have been dropped either at the library preparation or
554 bioinformatic analysis step if standard quality control procedures were applied. In our study we aimed to
555 evaluate the widest possible range of issues encountered during metabarcoding analysis and test the effect
556 of different sources of variation (including extreme ones) on the final community data. The post-hoc
557 corrections applied to the datasets represent the usual quality checkpoints during sample processing and
558 initial data screening. For example, contaminated samples or those with low sequence yield would usually
559 be removed, as well as samples with extremely low sequence number or OTU diversity. These simple
560 corrections help to remove the noise from the dataset, introduced by the rare tail (i.e., singletons and
561 doubletons, that are in most cases amplification, sequencing or contamination artefacts), while conserving
562 and emphasizing the core patterns in the communities.

563

564 **Conclusions**

565 The results of this study suggest that, overall, community metabarcoding is relatively robust to
566 the random effects of laboratory-based variation within established sample processing protocols and
567 largely confirm our initial hypotheses: (1) Metabarcoding results from all laboratories provided consistent
568 patterns of discrimination among four community samples for both DNA markers (COI and 18S rRNA),
569 despite considerable variation; (2) The two different HTS library protocols (dual index primers *versus*
570 fusion primers) did not significantly affect the community comparisons, even though they appeared to
571 have some differential effects on numbers and quality of sequence reads from the two markers; (3) We
572 identified several factors that introduced the greatest variability (preservation buffer, sample defrosting,
573 template concentration, DNA polymerase, PCR enhancer [BSA]); (4) We confirmed that standard post-
574 hoc data filtering steps (e.g., excluding samples with low sequence number or eliminating contaminating
575 sequences) were very effective at removing noise in the metabarcoding data introduced by laboratory

576 variation. The main findings and further considerations resulting from our study are summarized in the
577 provisional guidelines shown below (Text box 1).

Text box 1: Considerations for improved standardization and minimising effects of laboratory-based variation in metabarcoding community analyses.

Primer choice

18S rRNA: better detection of large-scale patterns, less sensitive to contamination

COI: better species resolution, higher sequence yield, less sensitive to freeze-thaw cycles

Library type

Dual index: higher sequence yield for the 18S rRNA gene, less susceptible to inhibitors,

Fusion: higher sequence yield for the COI gene, less prone to contamination

Sample handling and processing factors substantially affecting metabarcoding data variability

- Preservation buffer
- Type of DNA polymerase
- PCR enhancers

General recommendations

- Immediate processing or cold storage (ideally at -20°C or lower temperatures), whenever practical
- Preservation buffer type, temperature and storage time should be recorded and reported in the metabarcoding protocols and considered as covariates, whenever relevant
- Keep track of and report freeze-thaw cycles of DNA samples (avoid if possible)
- Divide DNA into multiple aliquots upon extraction
- Caution around using BSA and other PCR inhibitors, as these might affect variance and composition of investigated communities
- Technical replication at the PCR step is recommended to mitigate effects of within sample variability
- Consider appropriate data correction and de-noising for reducing effects from sequencing or contamination artefacts and better discerning core biological patterns
- Keep track and report all data corrections

Further considerations

- The effect of slightly different workflows may be more significant when discriminating assemblages with moderate overlap but some turnover (β diversity) in species assemblages. Therefore, bias rate imposed by methodological variation deserves further investigation at fine-scale levels for species diversity estimates.
- Due to large variation of sample types, primers, target taxa, differential laboratory access to reagents and many other factors, it is impractical to completely standardize PCR protocols across all metabarcoding studies. However, the maximum possible standardization is required in studies aimed at comparative biodiversity analyses.
- It is critical to clearly articulate methods in publications to enable a better understanding of the underlining causes of biological deviations or lack of reproducibility between studies.
- It is impossible to control the effects of all possible laboratory parameters in one study, therefore further comparative and cross-calibrations studies, at both large and small spatial scales, should be of great assistance for further identifying the key factors introducing biological variation and for improving standardization of metabarcoding protocols.

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588

589 **Author Contributions**

590 AZ, PG, CA, UvA, MB, MEC, JG, AAG, MH, GJI, SL, AS, TS, MS, JS, SAW and XP conceived the
591 study. AZ, CA, UvA, JG, MS, SAW and XP defined the design and collected the field samples. AZ, CA,
592 UvA, JB, MB, MEC, AC, ED, JG, AAG, MH, EH, SL, AS, TS, MS, SS, JS, VT, KW, MW and GZ
593 performed sample processing and library preparation. MB, MH, TS, MS and MW run the sequencing. PG
594 run bioinformatic analyses on the datasets. AZ performed statistical analyses and produced the first draft
595 of the manuscript. AZ, PG, CA, MEC, AC, GJI, SL, TS, SAW and XP revised the early versions of the
596 manuscript and contributed to writing. All authors contributed to revision of later versions and final
597 proof-reading.

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601 **References**

- 602 Ahn, J. H., Kim, B. Y., Song, J., & Weon, H. Y. (2012). Effects of PCR cycle number and DNA polymerase
603 type on the 16S rRNA gene pyrosequencing analysis of bacterial communities. *J Microbiol*, *50*(6),
604 1071-1074. doi:10.1007/s12275-012-2642-z
- 605 Alberdi, A., Aizpurua, O., Gilbert, M. T. P., & Bohmann, K. (2018). Scrutinizing key steps for reliable
606 metabarcoding of environmental samples. *9*(1), 134-147. doi:10.1111/2041-210x.12849
- 607 Anderson, M. J. (2001). A new method for non-parametric multivariate analysis of variance. *Austral*
608 *Ecology*, *26*, 32-46.
- 609 Anderson, M. J. (2017). Permutational Multivariate Analysis of Variance (PERMANOVA). In N.
610 Balakrishnan, T. Colton, B. Everitt, W. Piegorisch, F. Ruggeri, & J. L. Teugels (Eds.), *Wiley StatsRef:*
611 *Statistics Reference Online* (pp. 1-15).
- 612 Aylagas, E., Borja, Á., Irigoien, X., & Rodríguez-Ezpeleta, N. (2016). Benchmarking DNA Metabarcoding
613 for Biodiversity-Based Monitoring and Assessment. *Frontiers in Marine Science*, *3*, Article 96.
614 doi:10.3389/fmars.2016.00096
- 615 Aylagas, E., Borja, A., & Rodrigues-Ezpeleta, N. (2014). Environmental status assessment using DNA
616 metabarcoding: towards a genetic based marine biotic index (gAMBI). *PLoS One*, *9*(3), e90529.
617 doi:10.1371/journal.pone.0090529.
- 618 Baillet, B., Apothéoz-Perret-Gentil, L., Baričević, A., Chonova, T., Franc, A., Frigerio, J.-M., . . . Kahlert, M.
619 (2020). Diatom DNA metabarcoding for ecological assessment: Comparison among
620 bioinformatics pipelines used in six European countries reveals the need for standardization.
621 *Science of The Total Environment*, *745*, 140948.
622 doi:<https://doi.org/10.1016/j.scitotenv.2020.140948>
- 623 Baker, M. (2016). 1,500 scientists lift the lid on reproducibility. *Nature*, *533*, 452-454.
- 624 Binladen, J., Gilbert, M. T., Bollback, J. P., Panitz, F., Bendixen, C., Nielsen, R., & Willerslev, E. (2007). The
625 use of coded PCR primers enables high-throughput sequencing of multiple homolog
626 amplification products by 454 parallel sequencing. *PLoS One*, *2*(2), e197.
627 doi:10.1371/journal.pone.0000197
- 628 Bourlat, S. J., Haenel, Q., Finnman, J., & Leray, M. (2016). Preparation of Amplicon Libraries for
629 Metabarcoding of Marine Eukaryotes Using Illumina MiSeq: The Dual-PCR Method. In S. J.
630 Bourlat (Ed.), *Marine Genomics: Methods and Protocols* (pp. 197-207). New York, NY: Springer
631 New York.
- 632 Bowers, H., Pochon, X., von Ammon, U., Gemmell, N. J., Stanton, J.-A., Jeunen, G.-J., . . . Zaiko, A. (2021).
633 Towards optimization of eDNA/eRNA sampling technologies for marine biosecurity surveillance.
634 *Water*, *13*(8), 1113.
- 635 Braukmann, T. W. A., Ivanova, N. V., Prosser, S. W. J., Elbrecht, V., Steinke, D., Ratnasingham, S., . . .
636 Hebert, P. D. N. (2019). Metabarcoding a diverse arthropod mock community. *19*(3), 711-727.
637 doi:10.1111/1755-0998.13008
- 638 Brown, E. A., Chain, F. J. J., Zhan, A., MacIsaac, H. J., & Cristescu, M. E. (2016). Early detection of aquatic
639 invaders using metabarcoding reveals a high number of non-indigenous species in Canadian
640 ports. *Diversity and Distributions*, *22*, 1045-1059.
- 641 Carøe, C., & Bohmann, K. (2020). Tagsteady: A metabarcoding library preparation protocol to avoid false
642 assignment of sequences to samples. *20*(6), 1620-1631. doi:<https://doi.org/10.1111/1755-0998.13227>
- 643
- 644 Clarke, K. R., & Gorley, R. N. (2015). *PRIMER v7: User Manual/Tutorial*. PRIMER-E, Plymouth, UK.
- 645 Clarke, L. J., Trebilco, R., Walters, A., Polanowski, A. M., & Deagle, B. E. (2020). DNA-based diet analysis
646 of mesopelagic fish from the southern Kerguelen Axis. *Deep Sea Research Part II: Topical Studies*
647 *in Oceanography*, *174*. doi:<https://doi.org/10.1016/j.dsr2.2018.09.001>

648 Collins, R. A., Bakker, J., Wangensteen, O. S., Soto, A. Z., Corrigan, L., Sims, D. W., . . . Mariani, S. (2019).
649 Non-specific amplification compromises environmental DNA metabarcoding with COI. *10(11)*,
650 1985-2001. doi:10.1111/2041-210x.13276

651 Comtet, T., Sandionigi, A., Viard, F., & Casiraghi, M. (2015). DNA (meta)barcoding of biological invasions:
652 a powerful tool to elucidate invasion processes and help managing aliens. *Biological Invasions*,
653 *17*, 905-822.

654 Cordier, T., Alonso-Sáez, L., Apothéloz-Perret-Gentil, L., Aylagas, E., Bohan, D. A., Bouchez, A., . . .
655 Lanzén, A. (2020). Ecosystems monitoring powered by environmental genomics: A review of
656 current strategies with an implementation roadmap. doi:10.1111/mec.15472

657 Cristescu, M. E. (2014). From barcoding single individuals metabarcoding biological communities:
658 towards an integrative approach to the study of global biodiversity. *Trends in Ecology &*
659 *Evolution*, *29(10)*, 566-571.

660 Darling, J. A., & Mahon, A. R. (2011). From molecules to management: adopting DNA-based methods for
661 monitoring biological invasions in aquatic environments. *Environmental Research*, *111*, 978-988.

662 Darling, J. A., Pochon, X., Abbott, C. L., Inglis, G. J., & Zaiko, A. (2020). The risks of using molecular
663 biodiversity data for incidental detection of species of concern. *Diversity and Distributions*, in
664 press.

665 Djurhuus, A., Port, J., Closek, C. J., Yamahara, K. M., Romero-Maraccini, O., Walz, K. R., . . . Chavez, F. P.
666 (2017). Evaluation of Filtration and DNA Extraction Methods for Environmental DNA Biodiversity
667 Assessments across Multiple Trophic Levels. *4(314)*. doi:10.3389/fmars.2017.00314

668 Doi, H., Fukaya, K., Oka, S.-i., Sato, K., Kondoh, M., & Miya, M. (2019). Evaluation of detection
669 probabilities at the water-filtering and initial PCR steps in environmental DNA metabarcoding
670 using a multispecies site occupancy model. *Scientific Reports*, *9(1)*, 3581. doi:10.1038/s41598-
671 019-40233-1

672 Dowle, E., Pochon, X., Banks, J., Shearer, K., & Wood, S. A. (2016). Targeted gene enrichment and high
673 throughput sequencing for environmental biomonitoring: a case study using freshwater
674 macroinvertebrates. *Molecular Ecology Resources*, *16(5)*, 1240-1254.

675 Duran, R., & Cravo-Laureau, C. (2017). Protocols for Mudflat and Algal Mat In Situ Analysis. In T. J.
676 McGenity, K. N. Timmis, & B. Nogales (Eds.), *Hydrocarbon and Lipid Microbiology Protocols: Field*
677 *Studies* (pp. 305-317). Berlin, Heidelberg: Springer Berlin Heidelberg.

678 Edgar, R. C. (2016). UNOISE2: improved error-correction for Illumina 16S and ITS amplicon sequencing.
679 081257. doi:10.1101/081257 %J bioRxiv

680 Edgar, R. C., Haas, B. J., Clemente, J. C., Quince, C., & Knight, R. (2011). UCHIME improves sensitivity and
681 speed of chimera detection. *Bioinformatics*, *27*, 2194-2200.

682 Elbrecht, V., & Leese, F. (2015). Can DNA-based ecosystem assessments quantify species abundance?
683 Testing primer bias and biomass - sequence relationships with an innovative metabarcoding
684 protocol. *PLoS One*, *10(7)*, e0130324. doi:10.1371/journal.pone.0130324

685 Elbrecht, V., & Steinke, D. (2018). Scaling up DNA metabarcoding for freshwater macrozoobenthos
686 monitoring. *Freshwater Biology*, *64(2)*, 380-387.

687 Engelbrektsen, A., Kunin, V., Wrighton, K. C., Zvenigorodsky, N., Chen, F., Ochman, H., & Hugenholtz, P.
688 (2010). Experimental factors affecting PCR-based estimates of microbial species richness and
689 evenness. *International Society for Microbial Ecology Journal*, *4(5)*, 642-647.

690 Evans, N. T., Olds, B. P., Renshaw, M. A., Turner, C. R., Li, Y., Jerde, C. L., . . . Lodge, D. M. (2016).
691 Quantification of mesocosm fish and amphibian species diversity via environmental DNA
692 metabarcoding. *Molecular Ecology Resources*, *16(1)*, 29-41.

693 Ficetola, G. F., Pansu, J., Bonin, A., Coissac, E., Giguët-Covex, C., De Barba, M., . . . Taberlet, P. (2015).
694 Replication levels, false presences and the estimation of the presence/absence from eDNA
695 metabarcoding data. *Molecular Ecology Resources*, *15(3)*, 543-556.

696 Geller, J., Meyer, C., Parker, M., & Hawk, H. (2013). Redesign of PCR primers for mitochondrial
697 cytochrome c oxidase subunit I for marine invertebrates and application in all-taxa biotic
698 surveys. *Molecular Ecology*, *13*(5), 851-861.

699 Goldberg, C. S., Turner, C. R., Deiner, K., Klymus, K. E., Thomsen, F. P., Murhpy, M. A., . . . Taberlet, P.
700 (2016). Critical considerations for the application of environmental DNA methods to detect
701 aquatic species. *Methods in Ecology and Evolution*, *7*, 1299-1307.

702 Gomez-Silvan, C., Leung, M. H. Y., Grue, K. A., Kaur, R., Tong, X., Lee, P. K. H., & Andersen, G. L. (2018). A
703 comparison of methods used to unveil the genetic and metabolic pool in the built environment.
704 *Microbiome*, *6*(1), 71. doi:10.1186/s40168-018-0453-0

705 Gray, M. A., Pratte, Z. A., & Kellogg, C. A. (2013). Comparison of DNA preservation methods for
706 environmental bacterial community samples. *FEMS Microbiol Ecol*, *83*(2), 468-477.
707 doi:10.1111/1574-6941.12008

708 Hajibabaei, M., Shokralla, S., Zhou, X., Singer, G. A., & Baird, D. J. (2011). Environmental barcoding: a
709 next-generation sequencing approach for biomonitoring applications using river benthos. *PLoS*
710 *One*, *6*(4), e17497.

711 Hampton-Marcell, J. T., Frazier, A., Moormann, S. M., Owens, S. M., & Gilbert, J. A. (2017). Preparation
712 and Analysis of Metatranscriptomic Libraries in Petroleum Hydrocarbon Microbe Systems. In T.
713 J. McGenity, K. N. Timmis, & B. Nogales (Eds.), *Hydrocarbon and Lipid Microbiology Protocols:
714 Genetic, Genomic and System Analyses of Communities* (pp. 51-67). Berlin, Heidelberg: Springer
715 Berlin Heidelberg.

716 Jeunen, G.-J., Knapp, M., Spencer, H. G., Taylor, H. R., Lamare, M. D., Stat, M., . . . Gemmell, N. J. (2019).
717 Species-level biodiversity assessment using marine environmental DNA metabarcoding requires
718 protocol optimization and standardization. *9*(3), 1323-1335. doi:10.1002/ece3.4843

719 Judo, M. S., Wedel, A. B., & Wilson, C. (1998). Stimulation and suppression of PCR-mediated
720 recombination. *Nucleic Acids Research*, *26*(7), 1819-1825. doi:10.1093/nar/26.7.1819

721 Keeley, N. B., Wood, S. A., & Pochon, X. (2018). Development and preliminary validation of a multi-
722 trophic metabarcoding biotic index for monitoring benthic organic enrichment. *Ecological
723 Indicators*, *85*, 1044-1057.

724 Kreader, C. A. (1996). Relief of amplification inhibition in PCR with bovine serum albumin or T4 gene 32
725 protein. *62*(3), 1102-1106.

726 Lacoursière-Roussel, A., Howland, K., Normandeau, E., Grey, E. K., Archambault, P., Deiner, K., . . .
727 Bernatchez, L. (2018). eDNA metabarcoding as a new surveillance approach for coastal Arctic
728 biodiversity. *8*(16), 7763-7777. doi:10.1002/ece3.4213

729 Larson, E. R., Graham, B. M., Achury, R., Coon, J. J., Daniels, M. K., Gambrell, D. K., . . . Suarez, A. V.
730 (2020). From eDNA to citizen science: emerging tools for the early detection of invasive species.
731 *18*(4), 194-202. doi:10.1002/fee.2162

732 Lee, K. M., Adams, M., & Klassen, J. L. (2019). Evaluation of DESS as a storage medium for microbial
733 community analysis. *PeerJ*, *7*, e6414. doi:10.7717/peerj.6414

734 Lehtiniemi, M., Ojaveer, H., David, M., Galil, B., Gollasch, S., McKenzie, C., . . . Pederson, J. (2015). Dose
735 of truth - Monitoring marine non-indigenous species to serve legislative requirements. *Marine
736 Policy*, *54*, 26-35.

737 Leray, M., & Knowlton, N. (2016). Censusing marine eukaryotic diversity in the twenty-first century.
738 *Philos Trans R Soc Lond B Biol Sci*, *371*(1702), 20150331. doi:doi:10.1098/rstb.2015.0331

739 Leray, M., & Knowlton, N. (2017). Random sampling causes the low reproducibility of rare eukaryotic
740 OTUs in Illumina COI metabarcoding. *PeerJ*, *5*(e3006).

741 Leray, M., Yang, J. Y., Meyer, C. P., Mills, S. C., Agudelo, N., Ranwez, V., . . . Machida, R. J. (2013). A new
742 versatile primer set targeting a short fragment of the mitochondrial COI region for

743 metabarcoding metazoan diversity: application for characterizing coral reef fish gut contents.
744 *Frontiers in Zoology*, 10(1), 34. doi:10.1186/1742-9994-10-34

745 Marotz, C., Sharma, A., Humphrey, G., Gottel, N., Daum, C., Gilbert, J. A., . . . Knight, R. (2019). Triplicate
746 PCR reactions for 16S rRNA gene amplicon sequencing are unnecessary. *67*(1), 29-32.
747 doi:10.2144/btn-2018-0192

748 McCord, B., Pionzio, A., & Thompson, R. (2014). *Analysis of the effect of a variety of PCR inhibitors on the*
749 *amplification of DNA using real time PCR, melt curves and STR analysis*. Retrieved from
750 [https://nij.ojp.gov/library/publications/analysis-effect-variety-pcr-inhibitors-amplification-dna-](https://nij.ojp.gov/library/publications/analysis-effect-variety-pcr-inhibitors-amplification-dna-using-real-time-pcr)
751 [using-real-time-pcr](https://nij.ojp.gov/library/publications/analysis-effect-variety-pcr-inhibitors-amplification-dna-using-real-time-pcr)

752 McMurdie, P. J., & Holmes, S. (2013). Phyloseq: an R package for reproducible interactive analysis and
753 graphics of microbiome census data. *PLoS One*, 8, e61217.

754 Molnar, J. L., Gamboa, R. L., Revenga, C., & Spalding, M. D. (2008). Assessing the global threat of invasive
755 species to marine biodiversity. *Frontiers in Ecology and the Environment*, 6(9), 485-492.

756 Murray, D. C., Coghlan, M. L., & Bunce, M. (2015). From Benchtop to Desktop: Important Considerations
757 when Designing Amplicon Sequencing Workflows. *PLoS One*, 10(4), e0124671.
758 doi:10.1371/journal.pone.0124671

759 Nichols, R. V., Vollmers, C., Newsom, L. A., Wang, Y., Heintzman, P. D., Leighton, M., . . . Shapiro, B.
760 (2018). Minimizing polymerase biases in metabarcoding. *18*(5), 927-939. doi:10.1111/1755-
761 0998.12895

762 Nicholson, A., McIsaac, D., MacDonald, C., Gec, P., Mason, B. E., Rein, W., . . . Hanner, R. H. (2020). An
763 analysis of metadata reporting in freshwater environmental DNA research calls for the
764 development of best practice guidelines. *2*(3), 343-349. doi:10.1002/edn3.81

765 O'Donnell, J. L., Kelly, R. P., Lowell, N. C., & Port, J. A. (2016). Indexed PCR Primers Induce Template-
766 Specific Bias in Large-Scale DNA Sequencing Studies. *PLoS One*, 11(3), e0148698.
767 doi:10.1371/journal.pone.0148698

768 Opel, K. L., Chung, D., & McCord, B. R. (2010). A study of PCR inhibition mechanisms using real time PCR.
769 *J Forensic Sci*, 55(1), 25-33. doi:10.1111/j.1556-4029.2009.01245.x

770 Pearman, J. K., von Ammon, U., Laroche, O., Zaiko, A., Wood, S. A., Zubia, M., . . . Pochon, X. (2021).
771 Metabarcoding as a tool to enhance marine surveillance of nonindigenous species in tropical
772 harbors: A case study in Tahiti. *3*(1), 173-189. doi:<https://doi.org/10.1002/edn3.154>

773 Porter, T. M., & Hajibabaei, M. (2018). Scaling up: A guide to high-throughput genomic approaches for
774 biodiversity analysis. *Mol Ecol*, 27(2), 313-338. doi:10.1111/mec.14478

775 Quast, C., Pruesse, E., Yilmaz, P., Gerken, J., Schweer, T., Yarza, P., . . . Glockner, F. O. (2013). The SILVA
776 ribosomal RNA gene database project: improved data processing and web-based tools. *Nucleic*
777 *Acids Res*, 41(Database issue), D590-596. doi:10.1093/nar/gks1219

778 R-project. (2014). R: A language and environment for statistical computing. Retrieved from
779 <http://www.R-project.org>

780 Renshaw, M. A., Olds, B. P., Jerde, C. L., McVeigh, M. M., & Lodge, D. M. (2015). The room temperature
781 preservation of filtered environmental DNA samples and assimilation into a phenol-chloroform-
782 isoamyl alcohol DNA extraction. *Mol Ecol Resour*, 15(1), 168-176. doi:10.1111/1755-0998.12281

783 Schrader, C., Schielke, A., Ellerbroek, L., & Johne, R. (2012). PCR inhibitors – occurrence, properties and
784 removal. *113*(5), 1014-1026. doi:<https://doi.org/10.1111/j.1365-2672.2012.05384.x>

785 Simberloff, D., Martin, J.-L., Genovesi, P., Maris, V., Wardle, D., Aronson, J., . . . Vilà, M. (2013). Impacts
786 of biological invasions: what's what and the way forward. *Trends in Ecology & Evolution*, 28(1),
787 58-66. doi:10.1016/j.tree.2012.07.013

788 Stat, M., Huggett, M. J., Bernasconi, R., DiBattista, J. D., Berry, T. E., Newman, S. J., . . . Bunce, M. (2017).
789 Ecosystem biomonitoring with eDNA: metabarcoding across the tree of life in a tropical marine
790 environment. *Scientific Reports*, 7, 12240.

791 Taberlet, P., Coissac, E., Hajibabaei, M., & Rieseberg, L. H. (2012). Environmental DNA: a special issue on
792 DNA metabarcoding. *Molecular Ecology*, *21*, 1789-1793.

793 Tatangelo, V., Franzetti, A., Gandolfi, I., Bestetti, G., & Ambrosini, R. (2014). Effect of preservation
794 method on the assessment of bacterial community structure in soil and water samples. *FEMS*
795 *Microbiol Lett*, *356*(1), 32-38. doi:10.1111/1574-6968.12475

796 Valentini, A., Taberlet, P., Miaud, C., Civade, R., Herder, J., Thomsen, P. F., . . . Dejean, T. (2016). Next-
797 generation monitoring of aquatic biodiversity using environmental DNA metabarcoding. *Mol*
798 *Ecol*, *25*(4), 929-942. doi:10.1111/mec.13428

799 van der Loos, L. M., & Nijland, R. (2020). Biases in bulk: DNA metabarcoding of marine communities and
800 the methodology involved. *Molecular Ecology*. doi:10.1111/mec.15592

801 Vandamme, S. G., Griffiths, A. M., Taylor, S. A., Di Muri, C., Hankard, E. A., Towne, J. A., . . . Mariani, S.
802 (2016). Sushi barcoding in the UK: another kettle of fish. *PeerJ*, *4*, e1891. doi:10.7717/peerj.1891

803 von Ammon, U., Wood, S. A., Laroche, O., Zaiko, A., Tait, L., Lavery, S., . . . Pochon, X. (2018). Combining
804 morpho-taxonomy and metabarcoding enhances the detection of non-indigenous marine pests
805 in biofouling communities. *Scientific Reports*, *8*, 16290.

806 Woide, D., Zink, A., & Thalhammer, S. (2010). Technical Note: PCR Analysis of Minimum Target Amount
807 of Ancient DNA. *American Journal of Physical Anthropology*, *142*(2), 321-327.

808 Wood, S. A., Smith, K. F., Banks, J. C., Tremblay, L. A., Rhodes, L., Mountfort, D., . . . Pochon, X. (2013).
809 Molecular genetic tools for environmental monitoring of New Zealand's aquatic habitats, past,
810 present and the future. *New Zealand Journal of Marine and Freshwater Research*, *47*(1), 90-119.

811 Zaiko, A., Pochon, X., Garcia-Vazquez, E., Olenin, S., & Wood, S. A. (2018). Advantages and limitations of
812 environmental DNA/RNA tools for marine biosecurity: management and surveillance of non-
813 indigenous species. *Frontiers in Marine Science*, *5*(322).

814 Zaiko, A., Samuiloviene, A., Ardura, A., & Garcia-Vazquez, E. (2015). Metabarcoding approach for
815 nonindigenous species surveillance in marine coastal waters. *Marine Pollution Bulletin*, *100*(1),
816 53-59.

817 Zhan, A., Hulak, M., Sylvester, F., Huang, X., Adebayo, A. A., Abbott, C., . . . MacIsaac, H. J. (2013). High
818 sensitivity of 454 pyrosequencing for detection of rare species in aquatic communities. *Methods*
819 *in Ecology and Evolution*, *4*(6), 558-565.

820 Zizka, V. M. A., Elbrecht, V., Macher, J.-N., & Leese, F. (2019). Assessing the influence of sample tagging
821 and library preparation on DNA metabarcoding. *19*(4), 893-899.
822 doi:<https://doi.org/10.1111/1755-0998.13018>

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824 **Data accessibility statement**

825 The HTS data supporting the results presented in this manuscript were uploaded into CSIRO Data Access
826 Portal and is accessible online at <https://doi.org/10.25919/5j5x-0711>.

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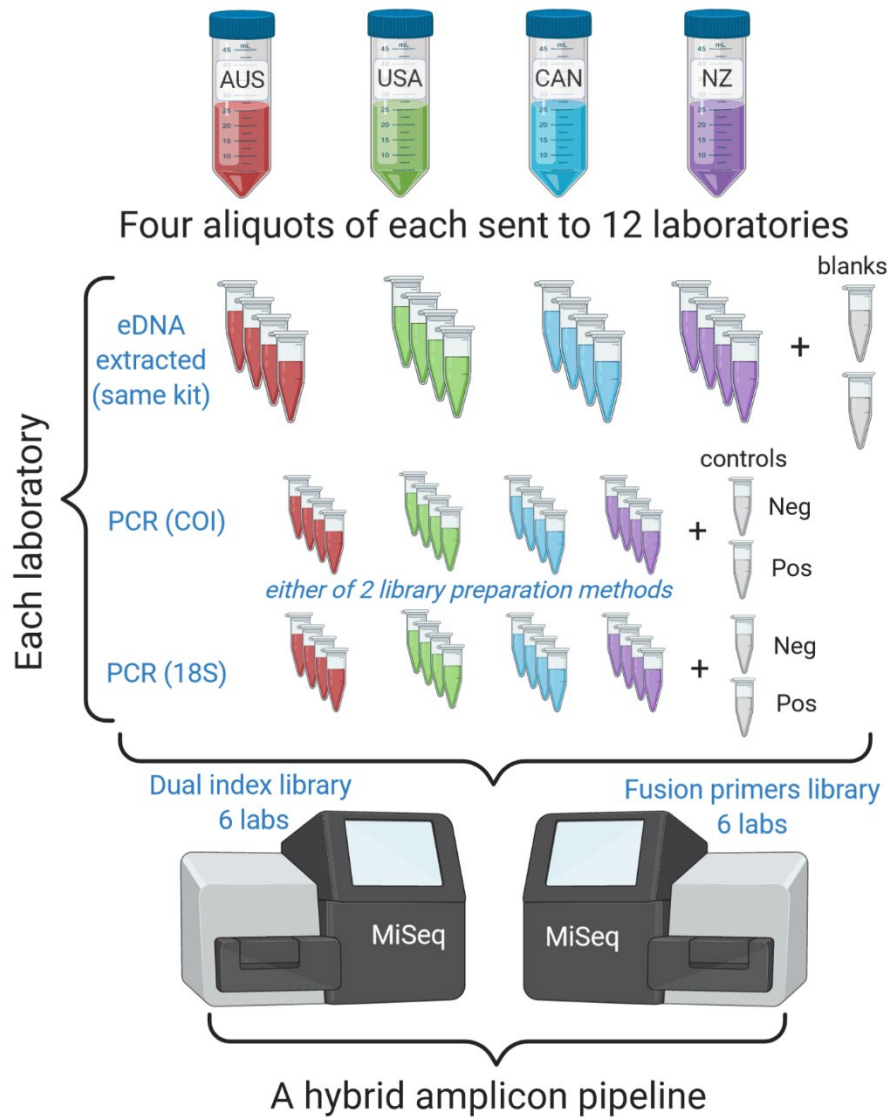
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Homogenized biofouling samples from 4 shipping hubs



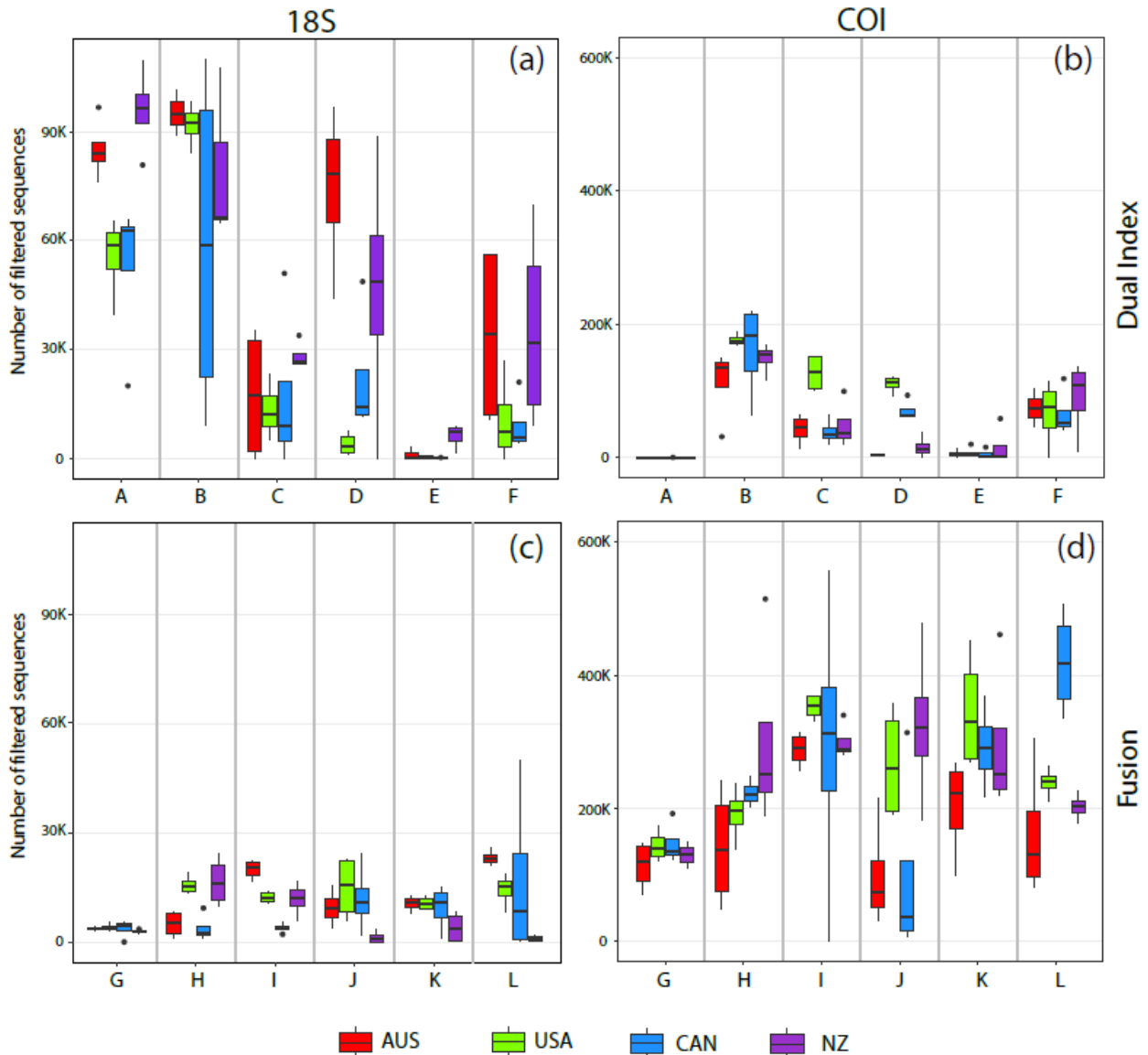
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836 Figure 1. Conceptual scheme of the experimental workflow (created with BioRender.com). Countries from
837 where the samples were sourced: AUS = Australia, USA = United States of America, CAN = Canada, NZ
838 = New Zealand. 18S = 18S ribosomal rRNA gene, COI = mitochondrial Cytochrome c Oxidase subunit 1.

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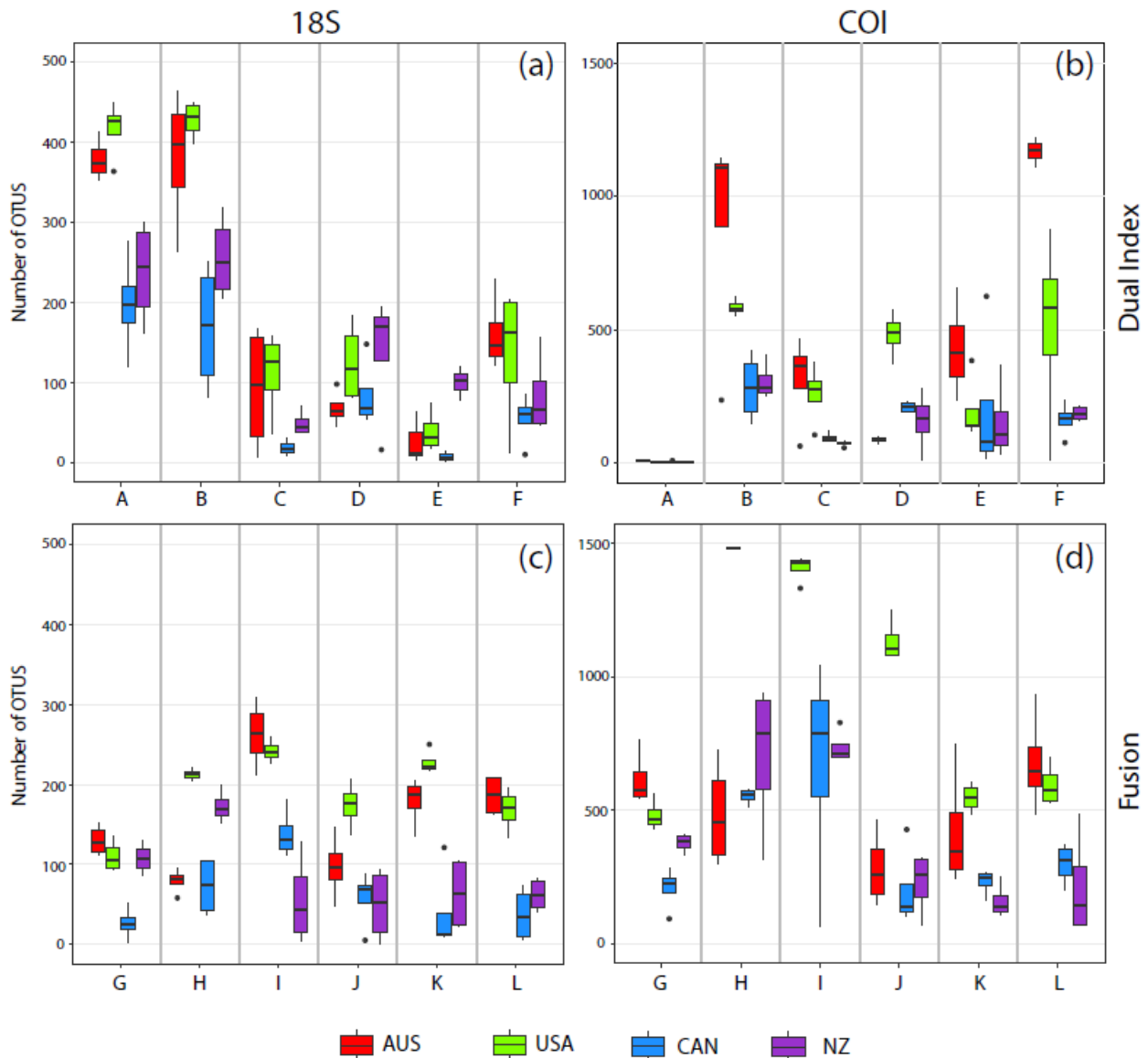
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845 Figure 2. Number of filtered 18S rRNA gene (a, c) and Cytochrome c Oxidase subunit 1 (b, d) sequences
846 obtained using the dual index (a, b) and fusion (c, d) methods. Each box plot represents the upper and
847 lower quartiles (edges), median (horizontal line), the maximum value of the data that is within 1.5 times
848 the interquartile range over the 75th percentile (the upper whisker), the minimum value of the data that is
849 within 1.5 times the interquartile range under the 25th percentile (the lower whisker) and the outliers
850 (black dots). The different colors represent the origin of the biofouling sample analyzed (AUS =
851 Australia, USA = United States of America, CAN = Canada, NZ = New Zealand). Note the difference in
852 the scale of y axis between markers.

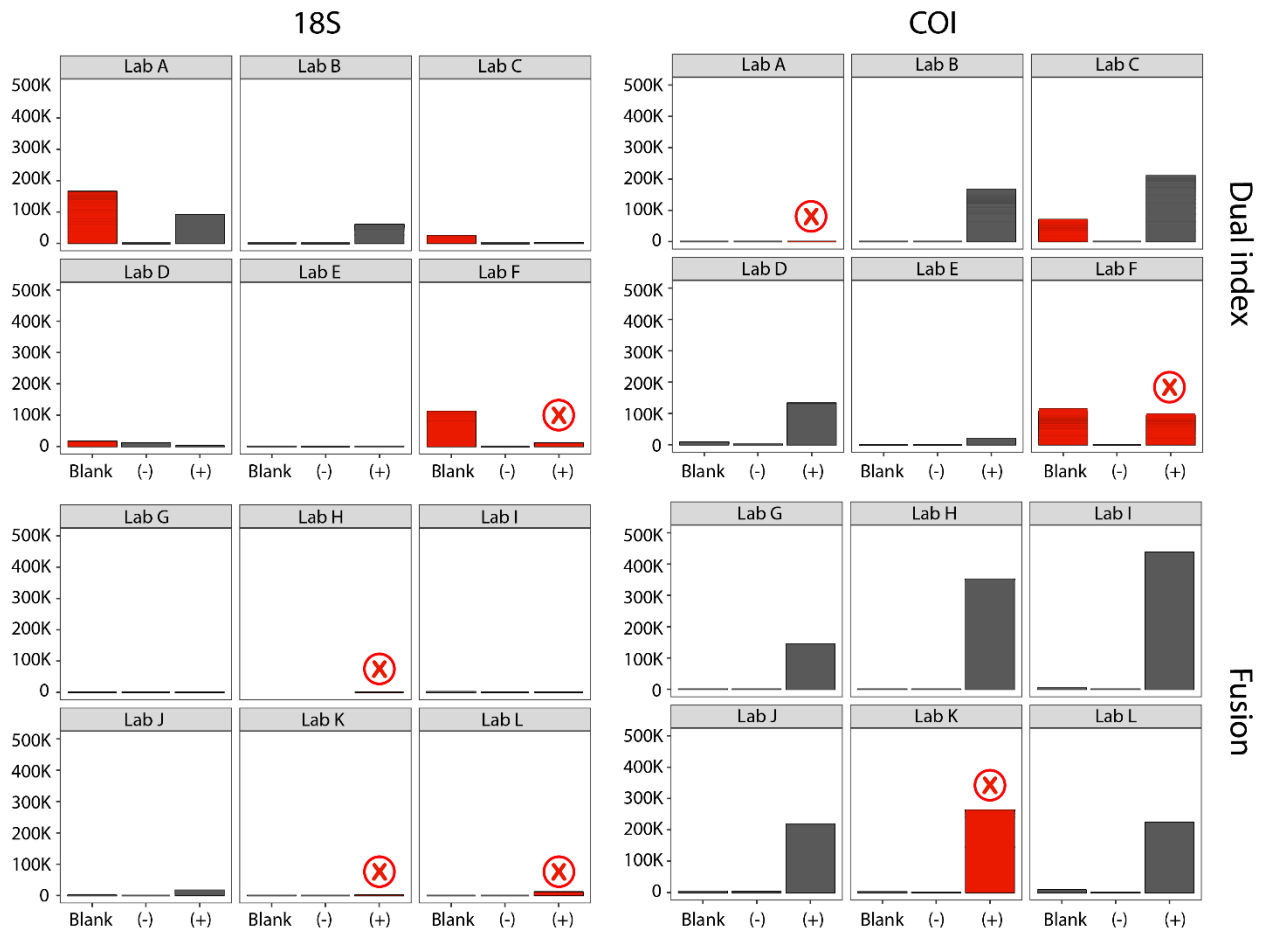


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854 Figure 3. Number of Operational Taxonomic Units (OTUs) obtained for each 18S rRNA gene (a, c) and
 855 Cytochrome c Oxidase subunit 1 (b, d) amplicons using the dual index (a, b) and fusion (c, d) methods.
 856 Each box plot represents the upper and lower quartiles (edges), median (horizontal line), the maximum
 857 value of the data that is within 1.5 times the interquartile range over the 75th percentile (the upper
 858 whisker), the minimum value of the data that is within 1.5 times the interquartile range under the 25th
 859 percentile (the lower whisker) and the outliers (black dots). The different colors represent the origin of the
 860 biofouling sample analyzed (AUS = Australia, USA = United States of America, CAN = Canada, NZ =
 861 New Zealand). Note the difference in the scale of y axis between markers.

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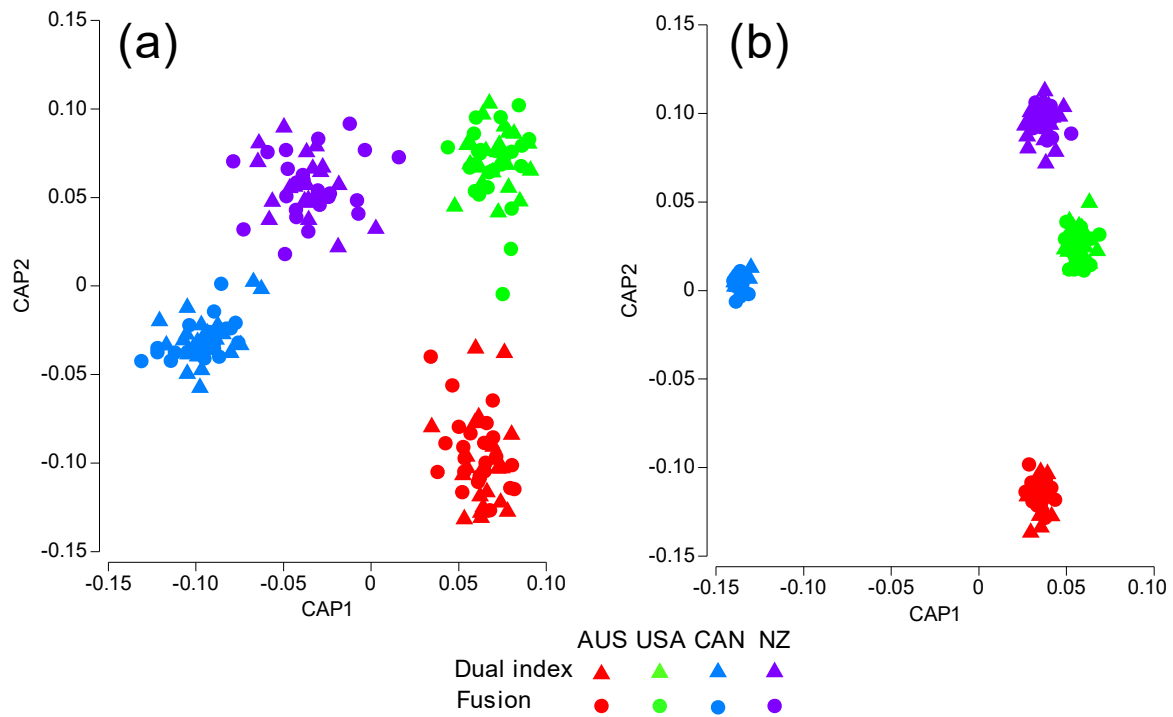
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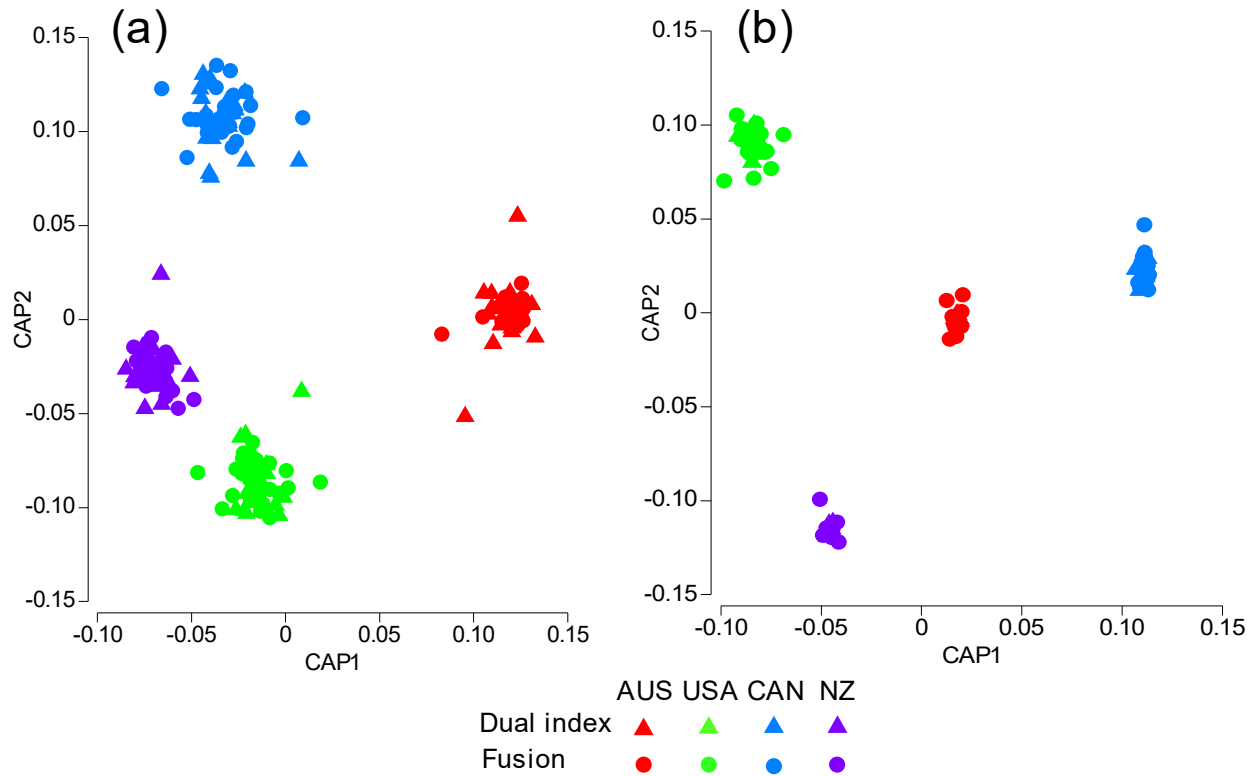
865 Figure 4. Number of filtered sequence reads recovered from the DNA extraction blanks (blank), negative
866 PCR controls (-) and positive PCR controls (+). Red bars indicate samples with considerable
867 contamination in the DNA extraction blanks; red crosses above red bars indicate positive PCR controls
868 that failed to recover the target taxa. 18S = 18S rRNA gene, COI = Cytochrome c Oxidase subunit 1.

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871 Figure 4. Canonical analysis of principal coordinates (CAP) plots based on Bray-Curtis dissimilarities of
 872 fourth root transformed 18S rRNA gene read abundance data using sample origin as a grouping factor: a)
 873 raw, uncorrected data; b) Corr1 data (samples with low sequence number (<1,000) excluded).

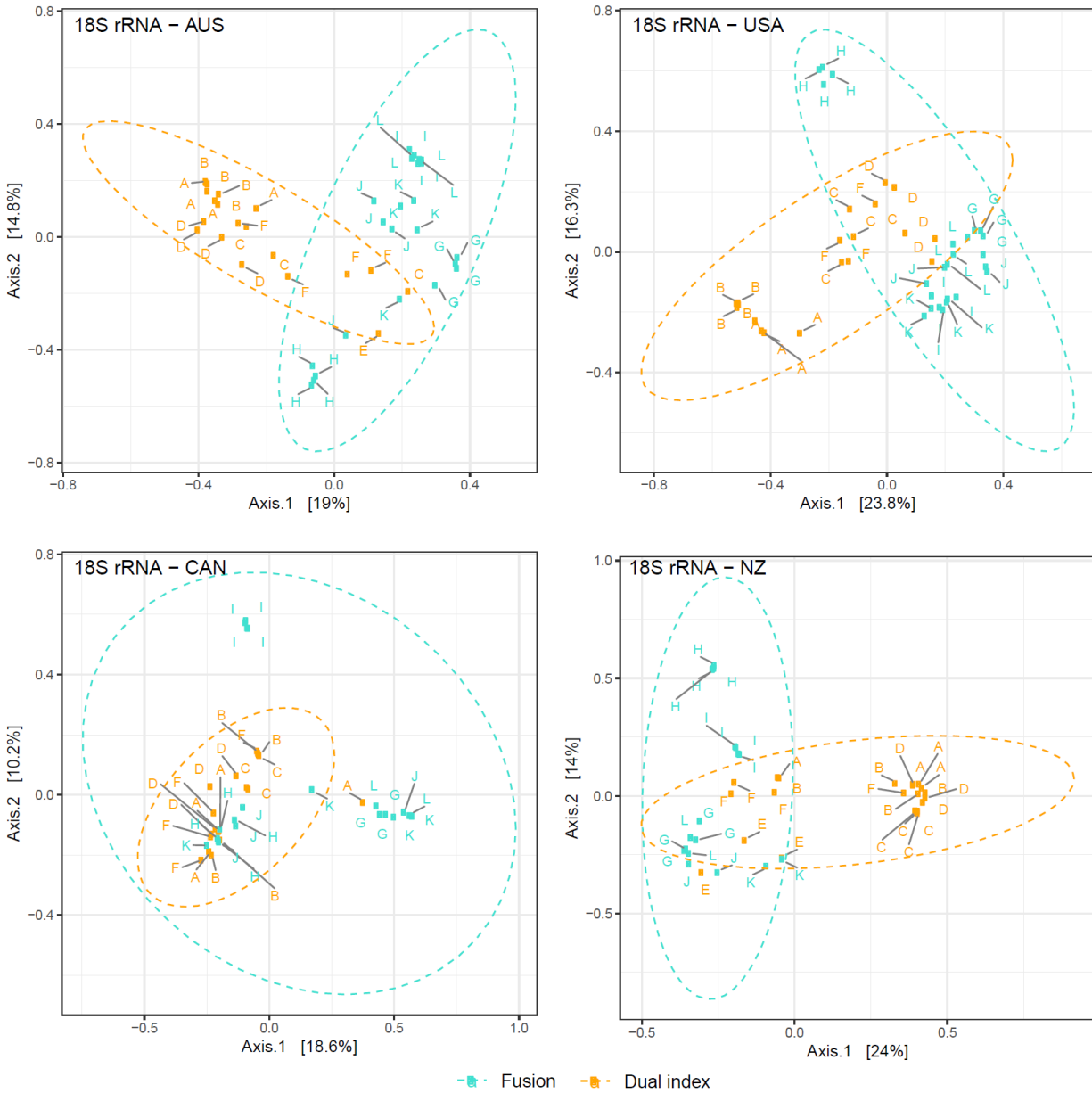
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877 Figure 5. Canonical analysis of principal coordinates (CAP) plots based on Bray-Curtis dissimilarities of
878 fourth root transformed mitochondrial Cytochrome c Oxidase subunit 1 (COI) read abundance data using
879 sample origin as a grouping factor: a) raw, uncorrected data; b) Corr3 data (OTUs found in the negative
880 controls removed from the corresponding subset of data). Triangles represent samples from dual index
881 sequencing library, circles – samples from fusion sequencing library.

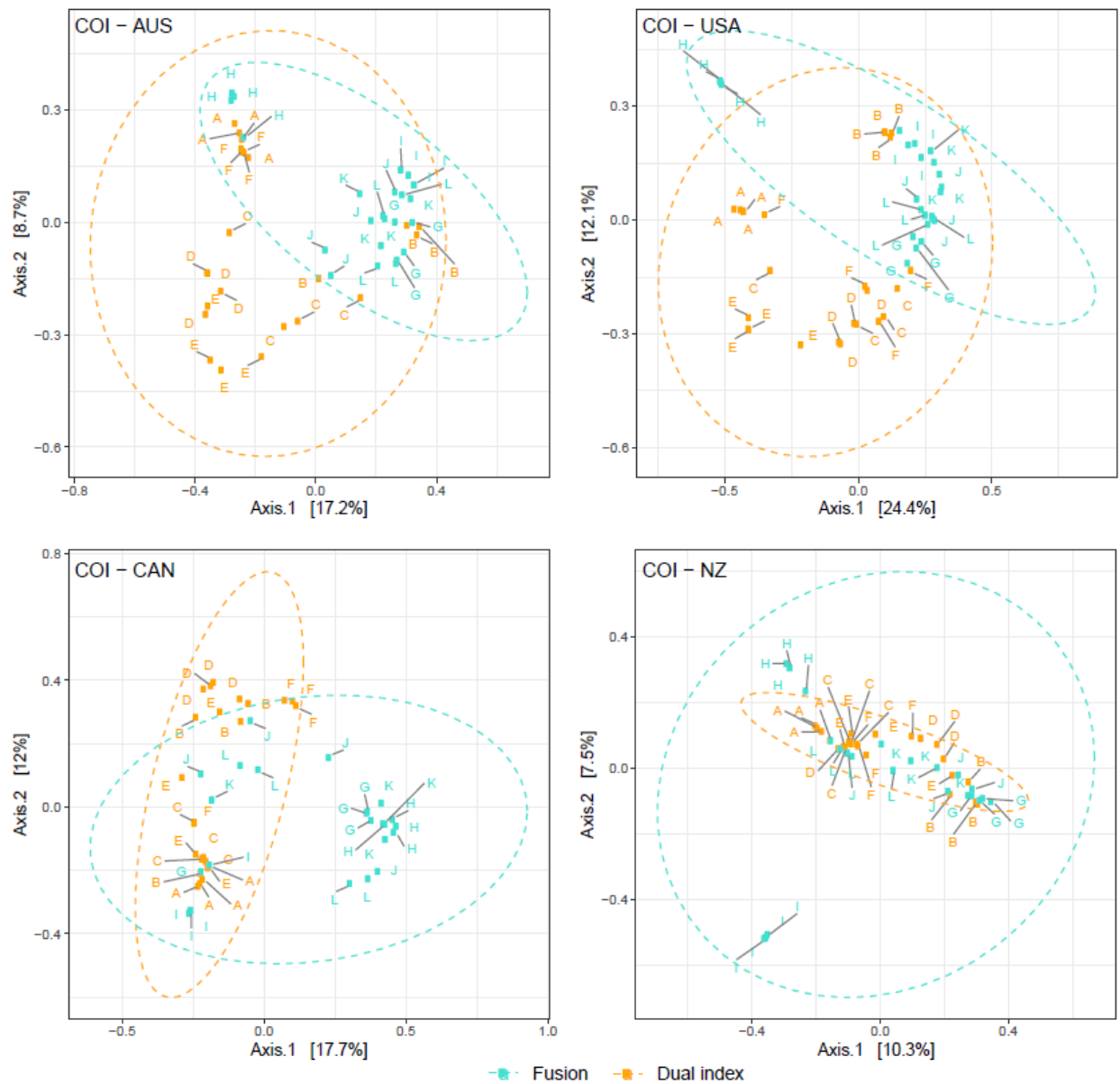
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884 Figure 6. Two-dimensional Principal Coordinates Analysis (PCoA) visualizations of 18S rRNA
 885 Operational Taxonomic Unit (OTU) diversity derived in the corrected datasets (Corr1) by different
 886 laboratories (laboratory labels noted for each datapoint) for samples originating from Australia (AUS),
 887 United States of America (USA), Canada (CAN) and New Zealand (NZ) samples.

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889

890 Figure 7. Two-dimensional Principal Coordinates Analysis (PCoA) visualizations of mitochondrial
 891 Cytochrome c Oxidase subunit 1 (COI) operational taxonomic unit (OTU) diversity derived in the
 892 corrected datasets (Corr3) by different laboratories (laboratory labels noted for each datapoint) for
 893 samples originating from Australia (AUS), United States of America (USA), Canada (CAN) and New
 894 Zealand (NZ) samples.

895

896 Table 1. Summary information on the samples of marine biofouling (starting material).

Sample origin	Sample code	Date	Source	Preservation
Hillarys Boat Harbour, Perth, Australia	AUS	8 February 2017	Settlement plate (scraped biomass, homogenized and frozen before preservation)	RNAlater™
Monterey Bay Harbour, California, USA	USA	1 May 2017	Settlement plate biomass (scraped biomass)	RNAlater™
Victoria Harbour, British Columbia, Canada	CAN	9 May 2017	Marina ropes (scraped biomass)	LifeGuard™
Waitematā Harbour, Auckland, North Island, New Zealand	NZ	20 April 2017	Marina pontoon (scraped biomass)	LifeGuard™

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899 Table 2. Summary of sample classification with raw and corrected 18S rRNA data from all laboratories.
 900 Canonical analysis of principal coordinates (CAP) analyses were carried out on fourth root transformed
 901 18S rRNA gene environmental DNA read abundance data obtained from 12 laboratories: Raw – not
 902 corrected; Corr1 – samples with low sequence number (<1,000) excluded; Corr2 – samples randomly
 903 rarefied to 10,000 sequence depth; Corr3 – Operational Taxonomic Units (OTUs) found in the negative
 904 controls removed from the corresponding subset of data; Corr4 – maximum read abundance of OTUs
 905 found in the negative controls subtracted from the corresponding subset of data; Corr5 – OTUs found in
 906 negative controls subtracted and samples with low sequence number (<1,000) excluded; Corr6 – OTUs
 907 found in negative controls subtracted and samples rarefied to 10,000 sequence depth. Shading indicates
 908 the best classification result of samples by origin.

18S rRNA gene	Raw	Corr1	Corr2	Corr3	Corr4	Corr5	Corr6
No. of samples	190	167	105	190	190	166	101
No. of OTUs	2,701	2,700	2,463	2,646	2,654	2,648	2,409
No. of sequences	4,860,418	4,855,461	1,050,000	2,956,370	4,548,860	4,543,847	1,010,000
Canonical correlation	0.961	0.997	0.998	0.981	0.953	0.974	0.997
P value	0.001	0.001	0.001	0.001	0.001	0.001	0.001
Correct classification %	88.42	92.81	91.43	87.89	87.89	90.36	91.09
AUS	89.36	90.91	100	87.23	91.49	90.7	96.43
USA	91.67	97.67	92.86	93.75	89.58	95.35	92.86
CAN	93.75	95	85.71	87.5	89.58	92.5	85.71
NZ	78.72	87.5	84	82.98	80.58	82.5	87.5

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916 Table 3. Summary of sample classification with raw and corrected mitochondrial Cytochrome c Oxidase
 917 subunit 1 (COI) data from all laboratories. Canonical analysis of principal coordinates (CAP) analyses
 918 were carried out on fourth root transformed mitochondrial COI eDNA read abundance data obtained from
 919 12 laboratories: Raw – not corrected; Corr1 – samples with low sequence number (<1,000) excluded;
 920 Corr2 – samples randomly rarefied to 10,000 sequence depth; Corr3 – Operational Taxonomic Units
 921 (OTUs) found in the negative controls removed from the corresponding subset of data; Corr4 – maximum
 922 read abundance of OTUs found in the negative controls subtracted from the corresponding subset of data;
 923 Corr5 – OTUs found in negative controls subtracted and samples with low sequence number (<1,000)
 924 excluded; Corr6 – OTUs found in negative controls subtracted and samples rarefied to 10,000 sequence
 925 depth. Shading indicates the best classification result of samples by origin.

COI Datasets	Raw	Corr1	Corr2	Corr3	Corr4	Corr5	Corr6
No. of samples	192	170	154	192	192	168	150
No. of OTUs	8,549	8,547	7,370	8,198	8,473	8,469	7,288
No. of sequences	27,866,837	27,865,889	1,540,000	13,049,764	27,286,419	27,284,006	1,500,000
Canonical correlation	0.98	0.993	0.994	0.999	0.996	0.997	0.993
p	0.001	0.001	0.001	0.001	0.001	0.001	0.001
Correct classification %	88.02	94.71	92.86	95.58	94.35	95.24	92.67
AUS	93.75	97.73	97.3	97.78	100	97.62	97.06
USA	91.67	97.67	92.5	93.48	95.46	97.67	95
CAN	83.33	92.68	89.74	91.3	93.18	95.12	92.11
NZ	83.33	90.48	92.11	100	88.64	90.48	86.84

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