1	Integrating species distribution models at forest planning level to develop
2	indicators for fast-growing plantations. A case study of <i>Eucalyptus globulus</i>
3	Labill. in Galicia (NW Spain)
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25 Abstract

26 Planting eucalypt stands in the Iberian Peninsula is controversial. In Galicia (NW Spain) the 27 area covered by eucalypts and the intensity of eucalypt timber harvesting are among the highest 28 in Europe. In this region, management plans for forest resources are intended for relatively 29 homogeneous territorial units, smaller than the province, called forest districts (ForDis). The 30 status of eucalypt plantations in ForDis is usually assumed to be highly variable, although this has not yet been quantified. Within this framework, the present study aimed to integrate an 31 32 existing Eucalyptus globulus distribution model for current conditions and two different climate change scenarios with current forest cover (including eucalypt plantations) to provide 33 34 quantitative and spatially-explicit information for identifying ForDis where eucalypt 35 plantations may become problematic. Four parameters (suitability, importance, potential 36 expansion and fragmentation) comprising 13 related indicators were calculated for eucalypt 37 plantations in each of the 19 ForDis in the region. The findings show that more than 44% of the 38 afforested area is covered by eucalypts in the province of A Coruña and in forest districts VI 39 and XVIII (in Lugo and Pontevedra, respectively). A huge continuous patch of eucalypt, of 40 more than 35000 ha, was detected across two ForDis (I and VI). The potentially high ecological 41 and financial risks associated with this vast area of monoculture would negatively affect the 42 sustainability and multifunctionality of these forest ecosystems. Analysis of the expected 43 suitability of areas for E. globulus in forthcoming climate change scenarios identified four 44 inland ForDis (IX, X in Lugo, XVI in Pontevedra and XI in Ourense) where an important 45 increase in cover by this species may occur if no action is taken. Moreover, these areas overlap 46 with the large ongoing expansion already observed for the more productive and frost-resistant 47 E. nitens. The findings also revealed edge-dominated native broadleaved forest patches in the habitat suitable for E. globulus in the provinces of A Coruña and Lugo. Although the current 48 49 level of occupancy of eucalypt within natural protected areas (NPAs) is moderate, the suitability 50 of these areas for eucalypt plantations will increase in the predicted climate change scenario,

51 possibly leading to aggravation of potential conflicts in these biodiversity hotspots. Regional

52 forest agencies should therefore increase monitoring efforts to prevent further impacts on native

53 forest and natural protected areas.

54 The proposed quantitative indicators and methodology used are intended to be useful for forest

55 planners and policy decision-makers in other regions where current and future conflicts related

- 56 to exotic tree plantations are expected.
- 57

58 Keywords

Tasmanian blue gum, forest planning, forest plantations, natural protected areas, species distributionmodels, climate change.

61 **1. Introduction**

62 Forests are increasingly recognized to be important for tackling many of the current great challenges as they provide many goods and services demanded by society (Freer-Smith et al., 63 64 2019). Although the global forest area decreased by 4.2% in the period 1990-2020, planted forest increased by around 72.4% in the same period (FAO, 2020). The increment is explained 65 by the response to the need to satisfy the increasing demand for timber resources, which has 66 67 mainly been supported by establishing plantations of fast-growing species belonging to genus 68 such as Pinus, Eucalyptus, Populus, Acacia and Tectona (FAO, 2010). These plantations also 69 play a dual role in helping to mitigate climate change (Freer-Smith et al., 2019) by i) acting as 70 efficient CO₂ sinks and *ii*) providing the raw material for producing thermal and electrical energy from forest biomass. Consequently, fast growing forest plantations will make an 71 72 important contribution to the inevitable transition of fossil-fuel-based economies to economies based on clean energy with a low carbon footprint. 73

74 Fast-growing plantations are often based on exotic species grown as monocultures with short 75 rotations, and greater control and planning efforts are thus required to guarantee sustainable 76 management and to preserve ecological processes and forest ecosystem biodiversity (Freer-77 Smith et al., 2019). Species distribution models (SDMs), which can be used to quantify 78 potentially suitable habitat for particular species (Guisan and Zimmermann, 2000), are powerful 79 tools for use in conservation planning, despite some limitations or weakness (McShea, 2014). 80 These models are particularly important when dealing with planted species for which the 81 current or future potential suitable habitat may already be covered by native forests or fall 82 within natural protected areas (NPAs) (López-Sánchez et al., 2021).

Within the framework of forest planning, SDMs can be used to predict future landscape
scenarios on the basis of projected climatic variables under different climate change scenarios.
This makes it possible to develop useful indicators that can be directly integrated in forest
planning documents enabling direct links with policy- and decision-making process (McShea,

87 2014). Such indicators can be direct or surrogate estimates of key features of forest ecosystems (such as integrity and resilience) and also enable assessment of temporal and spatial changes 88 (Ćosović et al., 2020). Overall, this type of information will be also useful for stakeholders, and 89 90 the integration of SDMs-derived outputs at forest planning scale should therefore contribute to 91 mitigating the so-called "assessment-planning gap". This gap refers to the difficulty in 92 constructing reliable conservation planning strategies due to non-applicable scientific findings 93 and the lack of true involvement of relevant stakeholders in the planning process (Villero et al., 94 2017).

95 The study area (Galicia) is an autonomous region in NW Spain with very suitable environmental 96 conditions for growing trees. The region is considered one of most productive forest areas in 97 Europe (García-Villabrille et al., 2014) and it is by far the most important timber-producing 98 region in Spain (Gómez-García, 2020). For more than a thousand years, until almost the middle 99 of the 20th century, a non-wooded forest landscape, with alternating shrubland, extensive 100 pastures and agricultural crops, clearly dominated in the region (Xunta de Galicia, 2018b). 101 Nevertheless, the increasing demand for timber for industrial purposes, together with the 102 suitable environmental conditions, led to wide scale afforestation from the middle of the 19th 103 century onwards, mainly with three highly productive species: Pinus pinaster Aiton, Pinus 104 radiata D. Don and Eucalyptus globulus Labill. In the last two decades, Eucalyptus nitens 105 (Deane and Maiden) Maiden has also been added to this group of highly productive species. 106 Plantations of these species, together with the abandonment of traditional agricultural activities, 107 act as the main drivers of the great forest landscape transformation in the last century in the 108 region (Marey-Pérez and Rodríguez-Vicente, 2009; Corbelle-Rico et al., 2015). Galicia was 109 thus transformed into an eminently forest region where 67.7% of the territory is currently 110 categorised as forest and almost a half as wooded land (Xunta de Galicia, 2018b). Plantations 111 of the above-mentioned major four species currently cover 57.46% of the total wooded area 112 (MAPA, 2019a) and play a key role in the Galician forestry sector, representing 92.27% of the average timber volume harvested in the period 2005-2016 in the region (7669591 m³) and almost a half of the same (48.66%) in the whole of Spain (MAPA, 2019b). Moreover, the existence of these plantations led to Galicia being one of the regions in Europe with the highest intensity of forest harvesting in the period 2000-2010 (Levers et al., 2014).

117 At landscape scale, the conversion of native forest or highly biodiverse natural vegetation into 118 forest plantations may have important consequences for native ecosystems, including habitat 119 loss and fragmentation, thus decreasing the size of remaining habitat patches and increasing 120 their isolation (Hanski, 1999). As a result, the probability of successful dispersal and 121 establishment of individuals and the persistence of populations in smaller and more isolated 122 habitat patches will be reduced, with negative impacts on local and regional species richness 123 (Harrison and Bruna, 1999; Teixido et al., 2010; Lomba et al, 2011). However, forest 124 plantations established on former agricultural or otherwise degraded land may provide 125 important opportunities for biodiversity conservation (Parrotta et al., 1997). Thus, the impact 126 of forest plantations on biodiversity will depend on the type of land use they replace (Pawson 127 et al., 2013) and also on the percentage of landscape covered. Analysis of landscape 128 configuration or spatial patterns is therefore very important in any study seeking to guarantee 129 ecosystem goods and services and particularly biodiversity conservation. This type of analysis 130 is usually carried out using a wide array of metrics describing the size, shape and isolation of 131 forest patches (e.g. McGarigal and Marks, 1994; Hargis et al., 1998).

The greatest expansion of eucalypt plantations occurred from the middle of the 20th century and was mediated by two major events: *i*) the beginning of the industrial production of pulp from eucalypt in Portugal in the 1950s (Bermudez et al., 2002) and *ii*) the installation of pulp mills in Galicia and in the nearby region of Asturias during the 1960s and 1970s (Riesco, 2004). Most of the plantations are of *E. globulus*, but since 1992, commercial plantations of *E. nitens* have also been established (Pérez-Cruzado, 2011). Both eucalypt species currently occupy an area of 345326 ha (MAPA, 2019a), which in the period 2005-2016 provided an annual average 139 timber harvest of 3787104 m³, representing 49.38% of the total volume harvested in the region 140 and 26.05% in the whole Spain (MAPA, 2019b). The rapid expansion of eucalypt species was mainly due to afforestation of non-wooded land (shrubland and abandoned agricultural land). 141 142 Nevertheless, the high profitability of commercial eucalypt plantations has led some forest owners to clear-cut patches of native forest for reforestation with E. globulus and E. nitens. 143 144 Moreover, some plantations were established in natural protected areas, aimed at protecting 145 biological diversity and the habitats that sustain the diversity. In these areas, uncontrolled 146 proliferation of exotic eucalypt plantations may jeopardise conservation goals.

The overall objective of this research was to provide indicators of the current and foreseeable future situation of eucalypt plantations in Galicia for use in formulating or revising forest planning documents (PORFs) at the forest district (ForDis) level. With this goal, we integrated an existing species distribution model (SDM) for *E. globulus* with spatial information about forest cover, current species distribution and timber harvesting rates. The case study focused on eucalypts plantations in Galicia due to the importance of these trees in Southern Europe.

153 **2. Materials and methods**

154 **2.1. Study area**

155 The study was carried out in Galicia (NW Iberian Peninsula) (Figure 1). The region occupies an area of 29575 km² and has a marked variation in elevation, ranging from 0-2127 m above 156 157 sea level, with a correspondingly variable climate. The region has a mild oceanic climate, with 158 an average annual temperature of 10–14 °C and average annual rainfall of 800–1500 mm, with 159 the lowest precipitation rates occurring during the summer months. The coastal area is 160 characterized by mild temperatures and abundant, uniformly distributed rainfall. The more continental inland area experiences more seasonally extreme temperatures. The predominant 161 162 soils are developed on granitic rock and acid schist, have a loam or sandy loam texture and are well drained. 163

Insert here Figure 1

164

165 **2.2. Forest land planning framework**

166 In Spain, responsibility for forest management was transferred from the Spanish government to 167 the autonomous governments in the middle of the 1980s. As a result, the autonomous 168 governments are currently almost the exclusive authorities in the forestry sector (MMA, 2002). 169 In the autonomous community of Galicia, the Galician Forest Plan (PFG) is the main forest 170 policy document and has been in force since 1992. The PFG provides the overall direction and 171 guidance for the entire afforested area in the medium and long term, in order to satisfy the 172 demand for resources by society and also to guarantee biodiversity conservation. The forest 173 planning documents, which develop PFG prescriptions, are generally known as "forest resource 174 management plans" (PORFs) and are intended for the territorial units designated as forest districts (ForDis). The 19 existing ForDis in Galicia cover large areas, although smaller than 175 176 the province, and are characterized by homogeneous geographic, socioeconomic, ecological 177 and cultural features (Marey Pérez et al., 2006; Xunta de Galicia, 2018a).

178 One of the major characteristics of the Galician forest sector is the type of forest ownership and 179 size. Thus, private forests comprise 97.3% of Galician forest land, of which 64.2% of plots of 180 mean size of 1.7 hectares are individually owned. This land is often divided into up to ten non-181 contiguous parcels, yielding a mean surface area of only 0.26 ha. Around 33.1% of private 182 forests are owned by communities and occupy an average area of 223 ha (Xunta de Galicia, 183 2018b). This scattered and very small private ownership is one of the primary hindrances to 184 further development of the sector (Robak et al., 2012), although great efforts have been made 185 during the last few decades to overcome this structural problem through land consolidation and 186 the creation of forest societies.

187 2.3. Sources of data

Four different sources of data were considered in this study: *i*) a map of the current distribution of eucalypt plantations; *ii*) a raster-based SDM of suitable habitat for *E. globulus*; *iii*) a map of types of forest cover; and *iv*) the average amount of timber harvested in 2018 and 2019 in each ForDis.

192 The current eucalypt distribution and area involved ($EU_{SNFI4.5}$) were obtained from the 2018 193 update of the Fourth Spanish National Forest Inventory for the most productive forest species 194 in northern Spain (hereinafter SNFI4.5) and its associated vectorial forest-cover map (scale 195 1:25000) (MAPA, 2019a). This map provides an accurate delineation of eucalypt-dominated 196 forests, with a minimum patch area of 2.5 ha (Alberdi et al., 2017).

In order to estimate the current and future suitable habitat for *E. globulus*, we used the rasterbased SDM developed for northern Spain by López-Sánchez et al., (2021). Three spatially continuous maps were obtained by applying the model developed to a list of current environmental variables (current suitable habitat) and to future climate variables reflecting two climate change scenarios, i.e. moderate (RCP 4.5) and pessimistic (RCP 8.5) for the 2050 time horizon (future suitable habitats).

203 To assess the distribution among ForDis of the types of forest cover that could be affected by a 204 hypothetical expansion of eucalypt in its suitable habitat, we used the 2018 version of the 205 CORINE Land Cover database (CLC2018). This database provides the land-related data at 206 20 m spatial resolution and considering a minimum mapping unit of 25 hectares (EEA, 2019). 207 The nomenclature of the vector data has 3 hierarchical levels. The classes in the first level are 208 artificial surfaces, agricultural areas, forests and semi-natural areas, wetlands and water bodies. 209 The second level has 15 classes and the third level, 44 subclasses. Although the more detailed 210 classes do not discriminate tree species, use of this database enabled us to differentiate different 211 types of forest cover (coniferous, broadleaved and mixed forest) and other types of land use. Statistics on timber harvested in 2018 and 2019 in each ForDis were recorded by the regional 212 213 government and are freely available online (Xunta de Galicia, 2020).

214 **2.4. Data analysis**

215 Data from the four sources of information were combined to produce metrics at two spatial 216 forest planning levels (ForDis and natural protected areas, NPAs) and for three time horizons

217 (current situation and two future scenarios under climate change projections) (Figure 2).

Insert here Figure 2

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219 2.4.1. Integrating current and potential suitable eucalypt areas with types of forest land cover

We first calculated metrics for different forest cover types from the following five layers: *i*) current and future suitable habitat for eucalypt obtained from the available SDM model (López-Sánchez et al., 2021); *ii*) current cover by broadleaved forest (code 311 of CLC2018); *iii*) current cover by coniferous and mixed forest (codes 312 and 313 of CLC2018); *iv*) current cover by other types of forest and semi-natural land (codes 321, 322, 323, 324, 333 and 334 of CLC2018); and *v*) current *E. globulus* (EUGL_{SNFI4.5}) and *E. nitens* (EUNI_{SNFI4.5}) cover, obtained from the SNFI4.5 (MAPA, 2019a). For the current situation, the land occupied by several forest cover categories was determined at two different spatial levels (ForDis and NPAs). For the spatial definition of the NPAs, we used the official, freely available vectorial mapping for these areas (MITECO, 2019). As result of this procedure, we determined the following four forest cover categories for each ForDis:

- Current area occupied by eucalypt plantations (EU_{SNFI4.5}), by summing EUGL_{SNFI4.5} plus
 EUNI_{SNFI4.5}.
- Current area occupied by native broadleaved forest (S_{NBF}) and the area included within the current eucalypt suitable habitat (SH_{NBF}). These were obtained on the basis of the difference between the current area of broadleaved forest (code 311 of CLC2018) and the current area occupied by eucalypt ($EU_{SNFI4.5}$) and acacia stands. This assumption appears valid, as the recent SNFI4.5 only inventoried 1217.6 ha of exotic broadleaved forest (*Acacia* sp.) within the current *E. globulus* suitable habitat in the whole study area.
- Current area occupied by coniferous and mixed forest (S_{CMF}) and the area included
 within the current eucalypt suitable habitat (SH_{CMF}). These were obtained by including
 the current area of conifer and mixed forest (codes 312 and 313 of CLC2018,
 respectively).
- Current area occupied by other natural and semi-natural land (S_{FSA}) and the area
 included in the current eucalypt suitable habitat (SH_{FSA}). These were obtained by
 considering the following CLC2018 codes: 321 (natural grassland), 322 (moor and
 heathland), 323 (sclerophyllous vegetation), 324 (transitional woodland-shrub), 333
 (sparsely vegetated areas) and 334 (burnt areas).

In addition, the areas occupied by the same forest cover categories were calculated by considering the future suitable habitat for *E. globulus*. For this purpose, the SDM developed for the species by López-Sánchez et al. (2021) was implemented for two different climate change scenarios predicted for 2050: moderate (RCP4.5) and pessimistic (RCP8.5). The "moderate scenario" (RCP 4.5) assumes a CO_2 concentration of 650 ppm and a temperature increase of 1.0–2.6 °C by 2100, whereas the "pessimistic scenario" (RCP 8.5) considers a CO_2 concentration of 1350 ppm and a temperature increase of 2.6–4.8°C by 2100.

256 2.4.2. Features analysed and indicators

Several issues concerning fast-growing forest plantations must be addressed in order to understand current and potential expansion and landscape configuration. These concerns are very important for guaranteeing ecosystem goods and services, and particularly biodiversity conservation. Within this framework, analysis of four factors in each ForDis was proposed in relation to the current and foreseeable future situation of eucalypt plantations: *i*) suitability, *ii*) importance, *iii*) potential expansion and *iv*) fragmentation (see Table 1 for a complete description of each).

We used the degree of suitability (DS) indicator to assess the suitability of the habitat for *E*. *globulus*. The current importance of eucalypt was assessed by both the degree of occupancy (DO) and the economic relevance (ER). In order to calculate these indicators, we considered the area occupied by both *E. globulus* and *E. nitens*, as timber harvesting data do not discriminate between these two species.

269 The potential for expansion of the species was assessed by considering the different forest cover 270 types currently existing within the suitable habitat for E. globulus. Thus, three different 271 potential expansion indicators were calculated: i) potential for replacement (PR), which 272 evaluates the potential replacement of native broadleaved forests by the exotic species E. 273 globulus; ii) potential for substitution (PS), which evaluates the potential substitution of other 274 non-native species (mainly *Pinus* spp. in the study area) by *E. globulus*; and *iii*) potential for 275 afforestation (PF), which evaluates the potential for establishment of new E. globulus forest in 276 natural and semi-natural non-wooded areas.

277 In addition, we used the FRAGSTATS 4.2 spatial analysis program (McGarigal *et al.*, 2016) to

278 quantify the degree of habitat fragmentation for eucalypt and native broadleaved cover. We

only analyzed these two types of forest cover as we wanted to compare the status of eucalypt plantations with the less altered native forests, which are of greater interest in term of biodiversity and other ecosystem services (Calviño-Cancela et al., 2012). We used several fragmentation indices related to patch size, core area and aggregation (Table 1). A 30 m-edge influence was applied for calculating the core area metrics (Laurance, 2000; Midha and Mathur, 2010).

Insert Table 1 here

286 **3. Results**

In this section we present the values of the thirteen indicators used for the two levels of spatial analysis (ForDis and NPAs) for the current situation and the 2050 time horizon under two climate change scenarios. Results of landscape fragmentation analysis for both ForDis and NPAs are also included.

291 **3.1.** Assessing the current situation within ForDis and NPAs

3.1.1. Eucalypt suitability and importance

293 The results obtained for DS, DO and ER indicators are shown in Table 3. According to these 294 results, a high proportion of the land in the province of A Coruña (94%) is suitable for growing 295 E. globulus, followed by Pontevedra (82%), Lugo (21%) and Ourense (with only 4%). These 296 results are consistent with the area currently occupied by the species in A Coruña, already more 297 than 200000 ha (Table 2), which is equivalent to 54% of degree of occupancy (DO) at province 298 level. At forest district level, the DO exceeded 45% in all ForDis and reached 65% in one of 299 them (ForDis I-Ferrol). The economic relevance (ER) indicates that eucalypt harvesting 300 accounts for more than 60% of the total volume harvesting in all the ForDis of the province, 301 reaching 88% in ForDis I-Ferrol (Table 3, Figure 3).

Insert Table 2 here

302

As a consequence of the very high suitability of the land for growing eucalypts, a total of 67% (3285 ha) of the whole area of eucalypt plantations included in NPAs in Galicia (4962 ha) is in the province of A Coruña. This area is mainly concentrated in ForDis I-Ferrol (Table 2). In A Coruña, the degree of suitability (DS) of NPAs for eucalypts is slightly lower (10% lower), than in the rest of the province, and the degree of occupancy (DO) is much lower (53.7% for the whole province compared with 29.5% for NPAs). The maximum DO value again corresponds to ForDis I-Ferrol (32.1%). 310 Eucalypt is globally moderately important in the province of Lugo according to the current 311 suitability and occupancy of the species (21% for both indicators). However, the highest 312 occupancy by eucalypt (DO) in the whole Galicia occurs in one seaboard forest district (ForDis 313 VI-A Mariña Lucense), with a DS of 90%, with 70% of the wooded surface occupied by 314 eucalypt (64000 ha) (Tables 2 and 3). Only one other ForDis in the province of Lugo has 315 significant suitability or occupancy of the species, the ForDis X-Terra Chá, which is close to 316 the coast. A guarter of the forest area in this ForDis is suitable for E. globulus and at present 317 21% of the wooded land is covered by eucalypt (Figure 3). Two ForDis (VII-A Fonsagrada-Os 318 Ancares and IX-Lugo-Sarria) with low eucalypt occupancy (DO of 1.6 and 7.1%) and 319 suitability (DS of 0.0 and 3.3%) have, however, higher ER values (19% and 22%, respectively) 320 than expected a priori. Globally, the presence of the species in NPAs in the province of Lugo 321 is very low (DO = 3%) and the suitability of the NPAs is 7% (Table 3). Nevertheless, DS values 322 of 58% and 19% and DO values of 31% and 7% were obtained for respectively the seaboard 323 ForDis VI-A Mariña Lucense and the neighbouring ForDis X-Terra Chá.

324 The province of Pontevedra is the third most important according to current area occupied by 325 eucalypt, with 52540 ha (Table 2). However, according to the suitability (82%) and degree of 326 occupancy (27%) at province level, the province ranks second. One ForDis (XIX-Caldas-O 327 Salnés) stands out, with 44% of its wooded area covered by eucalypt and an ER of 70%. The 328 DO values are between 10-30% and the ER between 32-45% for the remaining ForDis (Table 329 3, Figure 3). ForDis XVI- Deza-Tabeirós is similar to ForDis VII and IX in Lugo, with a low 330 presence of eucalypt (DO = 10%) but a relatively high economic relevance (ER = 32%). 331 Regarding the NPAs in this province, with the exception of ForDis XVI, with DS of 39%, the 332 DS is greater than 95% in the three remaining ForDis, although the DO is always lower than 333 10%. However, the indicators of potential expansion are always greater than 60% (Table 3).

334 Ourense is the only inland province of Galicia and this is the main reason why the suitability is

relatively high (DS = 24%) in only one ForDis (XI-O Ribeiro-Arenteiro), although the DO is

336 currently only 1%. In contrast to expectations considering the values of these indicators, and

337 similar to what is observed in ForDis VII, IX and XVI, the ER is 18% (Table 3, Figure 3). There

is no relevant presence of eucalypt in the NPAs in the province of Ourense (Table 3).

Insert Table 3, Figure 3 here

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340 **3.1.2.** Potential expansion of eucalypt

341 As a result of the almost total suitability of the land in the province of A Coruña for eucalypt, 342 the values of all of the potential expansion indicators are very high. The potential for 343 replacement (PR) values are higher than 90% for all ForDis, indicating that more than 90% of 344 the native forest is growing in suitable habitat for E. globulus. Similar values were obtained for 345 the potential for substitution (PS), indicating that most coniferous or mixed forest could also be 346 substituted by eucalypt. The percentage of natural and semi-natural non-wooded areas included 347 in the suitable habitat for *E. globulus* (potential for afforestation, PF) is high and greater than 75% in all ForDis. Analysis of the potential expansion indicators at the spatial level in NPAs 348 349 produced similar results, with average PR, PS and PF of 80-90% for the whole province and 350 values between 50-100% for the different ForDis (Table 3, Figure 3).

As for suitability and abundance, in the province of Lugo the potential for expansion was only relevant in the ForDis VI-A Mariña Lucense, with PR = 95%, PS = 75% and PF = 71% (Table 3). PR values of 65% and 19% and PS and PF values between 19 and 51% were obtained for the NPAs in the above-mentioned ForDis and its neighbouring ForDis X-Terra Chá.

Overall for the province of Pontevedra, the values of the indicators were 83% for PR, 90% for PS and 68% for PF; these values are very similar to those obtained for A Coruña. The potential rates of replacement, substitution and afforestation with eucalypt were also very high in the NPAs in the province of Pontevedra. In the inland province of Ourense only the ForDis XI-O Ribeiro-Arenterio displayed a significant potential for *E. globulus* expansion, with values of 360 19-32% for the expansion-related indicators (Table 3, Figure 3). Correspondingly, a significant

361 potential for replacement was only observed in this ForDis, with a value of 14%.

362 **3.1.3. Landscape fragmentation**

363 In the province of A Coruña, the mean patch size (MPS) is between three and fifteen times 364 greater for eucalypts than for native broadleaved forest. The greatest differences (15 and 8 times greater) occur in ForDis I-Ferrol and ForDis II-Bergantiños-Mariñas Coruñesas, respectively 365 366 (Table 4, Figure 3). The MPS of the native broadleaved forest is almost constant throughout the 367 province, with values of between 3 and 5 ha, whereas the values for eucalypt ranged from 15-368 38 ha. The largest patch index (LPI) indicates a huge continuous area occupied by eucalypt in 369 ForDis I, with a value of 26%, equivalent to almost 35000 ha of continuous eucalypt cover. 370 Remarkably, the largest area of native broadleaved forest patch (LPI = 1.6%), with 2150 ha of 371 continuous native forests, also occurs in this ForDis. Regarding the core area, the total core area 372 (TCA) of eucalypt plantations is 2-4 times greater than in native forest. The most surprising 373 finding was that the mean core area (MCA) is 3-19 times greater in eucalypt plantations. Indeed, 374 the highest values for eucalypts (35.8 and 17.2) and the lowest values for native forest (1.9 and 1.7) were observed in the above-mentioned ForDis, I and II respectively (Table 4). The 375 376 aggregation metrics indicate that in the whole province patch density (PD) is on average two 377 times greater in native forest than in eucalypt forest, but it is six times higher in ForDis I. The 378 nearest neighbor distance (ENN), as a consequence of greater fragmentation, is lower in native 379 forest (85 m) than in eucalypt plantations (120 m). The situation is more encouraging regarding 380 the fragmentation analysis of NPAs included within the eucalypt suitable habitat in the province 381 of A Coruña. Thus, MPS of native forest patches is 1-5 times greater than that of eucalypt 382 patches, although the maximum value of MPS in the five ForDis in the province does not exceed 383 9 hectares. The mean shape index (MSI) values are greater than in eucalypt patches, indicating 384 greater irregularity. Additionally, the total core area (TCA) and mean core area (MCA) are

higher in native forest patches than in eucalypt plantations patches, although the values of thelatter indicator are below 7 ha (Table 4).

387 In the province of Lugo, the worst scenario in terms of fragmentation of native forest in the 388 whole Galicia occurs in ForDis VI-A Mariña Lucense. In this ForDis, eucalypt MPS is twenty-389 five times greater than native forest (75 ha vs. 3 ha) with a huge LPI of 40%, which implies the 390 existence of continuous area occupied by eucalypt of more than 50000 ha (Figure 4). The MCA 391 in this ForDis is 62 ha for eucalypt and 2 ha for native broadleaved forest, which indicates that 392 most native forest patches are edge habitats (Figure 5). Within NPAs, similar MPS values were 393 found for eucalypt and native forest (around 3-4 ha), showing that the area occupied by 394 eucalypts is more fragmented but is still abundant, as indicated by the higher TCA of eucalypt 395 (1.5 times) than of native forest (Table 4). The MPS values in the other ForDis in the province 396 of Lugo are lower than 7.3 ha and are always lower than MPS values for native forest (up to 7.4 ha), and the situation in NPAs is similar. The TCA is also much higher in native forest than 397 398 eucalypt and the LPI comprised between 0.1 and 0.7% of the forest area, which implies 399 maximum areas of continuous occupation by eucalypt of up to 217 ha.

400 The inland province of Ourense does not have large areas of continuous cover by eucalypts and 401 only ForDis XI and XII have LPI values of 0.1%, which implies maximum continuous cover 402 by eucalypt of up to 25 ha. In Pontevedra, ForDis XVIII and XIX present highest LPI (2-2.2%) 403 which implies largest continuous patches of 1900-2400 ha and MPS reaches 20-25 ha with 7.9-404 6 ha for native forest, respectively. The MCA is 14.5-18.6 ha for eucalypt and 5.8-4.0 ha for 405 native forest and the PD is 0.47-0.9 in eucalypt and 1.7-2.9 in native forest indicating a high 406 degree of fragmentation of native forest. The remaining ForDis have the largest continuous 407 patches of eucalypt, of between 452-1080 ha, with greater LPI for native forest than eucalypt 408 (2.9-5.5 vs. 1.1-0.8).

Insert Table 4, Figure 4 here

409 **3.2.** Forecasting indicators in climate change scenarios

410 Projected indicator values for foreseeable climate change scenarios do not show important 411 changes for the province of A Coruña, as almost the whole province is already suitable for E. 412 globulus. However, some important changes are expected for two ForDis in each of the remaining three provinces. In Lugo, for the 2050 time horizon, the degree of suitability is 413 414 expected to increase by between 35 and 52% and the potential expansion indicators by between 415 33 and 62% for the non-mountainous inland ForDis (IX-Lugo-Sarria and X-Terra Chá) (Table 416 5). Globally, this province will increase the suitability for *E. globulus* by between 12 and 15% 417 for 2050 under moderate and pessimistic climate change scenarios, respectively. Future 418 projections also show two ForDis of the Ourense province with an important increase in 419 suitability for E. globulus (XI-O Ribeiro-Arenteiro and XII-Miño-Arnoia), which will increase 420 by between 57-70% and 30-60% for 2050 under moderate and pessimistic climate change 421 scenarios, respectively, with a similar increase in the expansion indicators (Table 5). In 422 Pontevedra, ForDis XVI- Deza-Tabeirós and ForDis XVII-O Condado-Paradanta are expected 423 to suffer increases in DS of 27-31% and 13-15% under the respective climate change scenarios. 424 The increases will also apply to the NPAs within each of the ForDis, so that a similar pattern to 425 that corresponding to the rest of the territory is expected for these natural areas (see Table 5).

Insert Table 5 here

427 **4. Discussion**

428 **4.1. Eucalypt suitability and importance**

A wide range of suitability and importance of current eucalypt plantations is observed across 429 the autonomous region of Galicia. The higher suitability for eucalypt in the seaboard provinces 430 431 without very mountainous areas (A Coruña and Pontevedra) and in ForDis VI-A Mariña 432 Lucense (in the coastal area of the province of Lugo) highlight the preference of the species in Galicia for a temperate climate (López-Sánchez et al., 2021). The high suitability in the 433 434 province of A Coruña is also accompanied by the importance of the species in this area, 435 reaching the highest percentage of the total wooded surface occupied by eucalypt in Northern 436 Spain and representing 75% of the mean annual volume harvested in the province. In the province of Pontevedra, both the suitability (between 63 and 95%, depending on ForDis) and 437 438 economic relevance (an average of 49%) indicators are slightly lower than in A Coruña. This 439 can be attributed to the fact that the degree of occupancy of pine plantations and mixed forest 440 in the province of Pontevedra is almost twice that in the province of A Coruña (MAPA, 2019a). 441 The province of Lugo includes the ForDis in Galicia with the highest percentage of eucalypt 442 cover over its wooded surface (ForDis VI- A Mariña Lucense). The environmental conditions 443 in the inland province of Ourense, with large mountain areas and important influence of 444 Mediterranean climate, are not very suitable for the species (López-Sánchez et al., 2021), and 445 therefore the values for the three indicators are very low.

Some surprising findings were obtained for the indicators in Four ForDis (VII-A Fonsagrada-Os Ancares, IX-Lugo-Sarria, X-O Riberiro-Arenteiro and XVI-Deza-Tabeirós). Thus, although the values for suitability and occupancy are very low in these areas, eucalypts account for between one fifth and one third of the total volume harvested, i.e. the economic relevance of eucalypt in those districts is rather high. There are two possible reasons for this finding: *i*) the very high productivity of *E. nitens* and *ii*) the "hidden" area occupied by eucalypt in these ForDis. During at least the past two decades, *E. nitens* has been planted by landowners in these 453 inland areas of Galicia (Pérez-Cruzado, 2011; González-García et al., 2015). The owners often 454 choose this species because it is more frost resistant than E. globulus (Prado and Barros, 1989). 455 Many of these plantations occupy very small areas (often less than 1 hectare) as a consequence 456 of the extremely high fragmentation of forest property in Galicia (Ónega-López et al., 2008). 457 These very small patches are not considered in the current forest cover maps because their 458 spatial resolution is lower than the patch size. Therefore, the economic importance indicator in 459 these ForDis becomes a suitable proxy for the existence of the "hidden" area occupied by 460 eucalypt. Moreover, due to the greater tolerance of E. nitens to lower temperatures, this species 461 is spreading through these ForDis towards inland Galicia.

462 In Galicia, a review of the PFG is currently ongoing, and once approved it will establish the 463 main guidelines for the Galician forestry sector for the next twenty years (Xunta de Galicia, 464 2018a). Forecasts for this time horizon in the PFG establish a maximum of 268000 ha for E. 465 globulus and 70000 ha for E. nitens (Xunta de Galicia, 2018b). Our results show that the current 466 area occupied by E. globulus exceeds this amount, and the target area forecast for E. nitens have 467 probably already been reached. These findings suggest that measures should be taken to reduce the abundance of eucalypts, although this does not necessarily imply a reduction in the 468 469 economic relevance of the species. A reduction in the area occupied can be accompanied by 470 increasing plantation productivity through the application of adequate silvicultural practices by 471 selecting more productive forest land for establishing new plantations and/or applying the 472 results of genetic improvement programs (Binkley et al., 2017). These measures can be also 473 complemented by the conversion of mixed eucalypt stands to pure pine or native broadleaved 474 forest, as already encouraged by the forest authorities through financial aids (DOG, 2019) and 475 even by forbidding replacement of current pine stands by eucalypt plantations.

Although eucalypt is not very common in NPAs, reduction of its presence in these areas is
desirable (Teixido et al., 2010). Regional regulations forbid the establishment of new eucalypt
plantations in sites included into the Natura 2000 network (DOG, 2014). Nevertheless,

plantations established prior to 2014 can continue to be managed in successive rotations (DOG,
2014). Most of the plantations were already present in most of the Natura 2000 sites whole the
Iberian Peninsula when these were designated as protected areas, as pointed out by Deus et al.
(2018).

483 **4.2. Potential expansion of eucalypt**

484 Forest districts within provinces where the suitability for *E. globulus* is currently high (A 485 Coruña and Pontevedra) also have high values for potential expansion (replacement, 486 substitution and afforestation). The large increase in suitability of this species under climate 487 change scenarios also indicates large increases in potential expansion in the future in six ForDis: 488 IX and X (Lugo), XI and XII (Ourense) and XVI and XVII (Pontevedra). Moreover, the 489 eucalypt-suitable habitat in the study area has already increased due to the introduction and 490 expansion of E. nitens (González-García et al., 2015). This species, which grows adequately in 491 the current E. globulus suitable habitat, has several features that make it preferred by forest 492 owners: *i*) it is frost tolerant (it tolerates temperatures as low as -10 °C; Prado and Barros, 1989); 493 *ii*) it shows a greater energy potential in the various tree biomass components compared to *E*. 494 globulus when forest waste is used for biomass production (Pérez et al., 2006); and iii) it 495 exhibits high growth rates (MAI in total volume with bark up to 46 m³/ha/year in the best sites; 496 Diéguez-Aranda et al. (2012), even greater than E. globulus (32-36 m³/ha/year); García-Villabrille, 2015; Viera et al., 2016). 497

In general, in Galicia, eucalypt plantations can be planted in former agricultural land and in shrublands (i.e. afforestation), or in former pine plantations (i.e. substitution). Nevertheless, plantations in former native forest (i.e. replacement) are forbidden by existing forest law (DOG, 2012). Nevertheless, this law is inadequately monitored and replacements do occur in certain areas.

Substitution of previous maritime and radiata pine plantations by eucalypts has occurred for
more than one decade, mainly in inland areas in the provinces of A Coruña and Lugo (González-

García et al., 2015). The species most commonly used is *E. nitens* (López-Sánchez et al., 2021),
which is even substituting previous *E. globulus* plantations in areas affected by *Gonipterus platensis* (> 350 m.a.s.l and shallow soils) as it is less affected by this pest (Martín Gil et al.,
2018; Gonçalves et al., 2019). A similar substitution process for conifer species has occurred
in Portugal (Mendes, 2007; Vaz et al., 2019).

510 As the forest authorities generally have insufficient resources for monitoring forestland, some 511 forest owners have replaced clear-cut native forest with new eucalypt plantations. To protect 512 and enhance native forests, the regional government has recently established an official 513 database of native broadleaved forest covering areas greater than 15 ha, so that owners will be 514 eligible for tax benefits and will be given priority for receiving public grants for sustainable 515 management and conservation (DOG, 2020). Nevertheless, this measure may be insufficient in 516 view of the high potential of replacement of E. globulus in forest districts in coastal provinces 517 and E. nitens in inland areas, which could lead to aggravation of the potential conflicts 518 concerning the natural forests. In particular, greater effort must be made to monitor eucalypt 519 expansion in some mountainous inland areas with large areas of native forest and where the economic relevance of eucalyptus reveals a significant presence of *E. nitens* in very small plots 520 521 (i.e. VII- A Fonsagrada-Os Ancares).

In summary, four major challenges arise concerning eucalypt expansion: *i*) high future values of potential expansion indicators for *E. globulus*; *ii*) current strong expansion of *E. nitens* (González-García et al., 2015); *iii*) increasing demand for raw material for energy production (e.g. a new biomass power plant consuming around 500000 Mg of forest biomass has recently been installed in Galicia (Rico, 2018)); and *iv*) foreseeable demand for raw material for new bioproducts, such as fibre for the textile industry (e.g. Xu et al., 2017; Yañez-S et al., 2018).

528 Indeed, eucalypt plantations represent one of the best ways for forest owners in Galicia to obtain

529 profits from their small, highly scattered plots (Díaz-Balteiro et al., 2009; Arenas et al., 2019).

530 Nonetheless, in view of the major challenges mentioned above, regulations for establishing new

eucalypt plantations are recommended. Specifically, decision-makers should be urged to
control substitution of previous coniferous plantations and to increase efforts involved in
monitoring native broadleaved forests to prevent their replacement.

534 **4.3. Landscape fragmentation**

The largest patch index (LPI) values indicated huge continuous areas of eucalypt in ForDis I-Ferrol and VI-A Mariña Lucense, of respectively 35000 and 50000 ha. Large LPI values were also observed in other ForDis within the province of A Coruña (range from 1470 to 10179 ha) and in the ForDis in Pontevedra (range from 429 ha to 2435 ha).

539 Very large areas of continuous eucalypt plantations should be avoided due to high biotic, abiotic 540 and financial risks (Freer-Smith et al., 2019) and also because this can exacerbate the negative 541 impacts of the species (Montero de Burgos, 1990; Veiras and Soto, 2011). Moreover, the biotic 542 and abiotic risks do not act independently, and the impacts may accumulate (Jactel et al., 2019). 543 From the point of view of forest land planning, encouraging owners to establish new plantations 544 with complementary tree species at stand level, in order to achieve mosaics of different forest 545 types at landscape level in the ForDis with such large LPI is strongly recommended (Freer-546 Smith et al., 2019).

547 Fragmentation analysis also revealed extreme fragmentation of native broadleaved forest in the 548 provinces of Lugo and A Coruña with very small mean patch size (4.1 ha, as average) and mean 549 core area (2.6 ha). Extreme fragmentation of native broadleaved forest and the above-mentioned 550 formations of large and continuous eucalypt plantations both occur as a direct result of the rapid 551 expansion of eucalypt throughout the territory and the replacement of patches of native forest 552 during the last few decades.

Replacement of native forest patches by eucalypt has led to the removal of stands that are much more diverse than eucalypt plantations (e.g. Bas López et al., 2018; Calviño-Cancela et al., 2012) and that can act as corridors connecting areas of high diversity. As a consequence of the small patch to small patch species replacement (consequence of the very small and highly 557 fragmented private forestland), remaining native broadleaved forests in seaboard areas under 558 strong pressure from eucalypt plantations are gradually reduced to form smaller and more fragmented and scattered patches (micro-patches). Patches with interior habitat are scarce and 559 560 almost all patches become edge habitat (mean core area 1.7-1.9 ha) thus greatly influencing the 561 forest matrix and strongly limiting the presence of sensitive-to-edge species (Harper et al. 2005: 562 Echeverría et al. 2007). Thus, different fragmentation studies in temperate forest have 563 established a threshold area of 2-5 ha for the initiation of interior forest (e.g. Levenson, 1981; 564 Ranney et al. 1981; Palik and Murphy 1990). The high level of fragmentation of native forest 565 is an important threat to biodiversity (SCBD, 2005) as it decreases landscape connectivity, and 566 patches become more isolated and increasingly affected by edge effects (Murcia, 1995). 567 Furthermore, these impacts also reduce the biotic integrity of ecosystems (Kroner et al., 2016) 568 and their ability to recover from disturbances (resilience) (Bregman et al., 2014). Moreover, in 569 the medium term, the changes in patterns of habitat fragmentation and connection will probably 570 have at least as large an impact on forest ecosystems as that caused by climate change (e.g. 571 Sinclair et al., 2010).

572 **5. Conclusions**

573 The so-called "management plans for forest resources" (PORFs) are used to develop the 574 regional forestry plan (PFG) in Galicia. To accurately design and implement future policies 575 regarding the allocation of forest use at forest district level, policy-makers (i.e. politicians and 576 land or forest planners) need indicators at this spatial operational level. The study findings 577 provide quantitative and spatially explicit information identifying where eucalypt plantations 578 are becoming problematic and could be used to correct and prevent future uncontrolled 579 expansion of the species. Thus, the threshold value of 45% for the current degree of occupancy 580 (DO) has already been exceeded in all ForDis in the province of A Coruña and ForDis VI and 581 XIX in Lugo and Pontevedra, respectively. Considering the expected suitability of E. globulus 582 in forthcoming climate change scenarios, we identified four inland ForDis (IX, X in Lugo, XVI 583 in Pontevedra and XI in Ourense) where important future increments are forecast. Moreover, 584 those areas overlap with the ongoing expansion already observed for E. nitens, and they are 585 therefore key ForDis for monitoring and controlling eucalypt expansion towards inland Galicia. 586 The huge continuous patches of eucalypt in seaboard ForDis in Galicia are subject to high 587 ecological and financial risks, which may negatively impact forest ecosystem sustainability and 588 multifunctionality. The study findings also identified edge-dominated native broadleaved forest 589 patches in the E. globulus suitable habitat in the provinces of A Coruña and Lugo. The current 590 degree of occupancy of eucalypt within NPAs is moderate and lower than outside of these areas. 591 Nevertheless, the suitability of habitat for eucalypt plantations will increase in the foreseeable 592 climate change scenario, which may lead to aggravation of the potential conflicts concerning 593 those highly biodiversity areas. It must also be noted that other considerations such as 594 regulations to control the species, socioeconomic and market demands changes and foreseeable 595 future increases in forests threats (including the appearance of new pests and diseases) will 596 probably play a very important role in limiting, reversing or even encouraging expansion of the

- 597 species. Thus, monitoring efforts by regional forest administrations should be increased to
- 598 prevent further impacts on native forest and natural protected areas.

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603 **7. References**

604	Alberdi, I., Cañellas, I., Vallejo Bombín, R., 2017. The Spanish National Forest Inventory:
605	history, development, challenges and perspectives. Pesq. flor. bras., Colombo, 37 (91),
606	361-368.

- Arenas, S., Rodríguez-Soalleiro, R., Díaz-Balteiro, L., 2019. Turno óptimo de *Eucalyptus nitens* en Galicia introduciendo la fiscalidad en el análisis. XII Congreso da Asociación
 Española de Economía Agraria, Lugo (Spain), 04-06 September 2019.
- Bas López, S., Guitián Rivera, J., Sobral, M., 2018. Biodiversidad en plantaciones de eucalipto
 y en robledales del sur de Galicia: plantas y aves. Nova Acta Científica Compostelana
 (Bioloxía) 25, 71-81.
- Binkley, D., Campoe, O., Alvares, C.A., Carneiro, R.L., Cegatta, I., Stape, J.L., 2017. The
 interactions of climate, spacing and genetics on clonal Eucalyptus plantations across
 Brazil and Uruguay. For. Ecol. Manag. 405, 271-283.
 https://doi.org/10.1016/j.foreco.2017.09.050
- 617 Bermudez, J.D., Touza, M.C., Sanz, F., 2002. Manual de la madera de eucalipto blanco.
 618 Fundación o Fomento da Calidade Industrial e Desenvolvemento Tecnolóxico de
 619 Galicia. Ourense.
- Bregman, T., Sekerciouglu, C.H., Tobias, J., 2014. Global patterns and predictors of bird
 species responses to forest fragmentation: Implications for ecosystem function and
 conservation. Biological Conservation 169(1), 372-383.
 https://doi.org/10.1016/j.biocon.2013.11.024
- 624 Calviño-Cancela, M., Rubido-Bará, M., Van Etenn, E., 2012. Do eucalypt plantations provide
 625 habitat for native forest biodiversity? Forest Ecology and Management 270, 153-162.

- Corbelle-Rico, E., Butsic, V., Enríquez-García, M.J., Radeloff, V.C., 2015. Technology or
 policy? Drivers of land cover change in northwestern Spain before and after the
 accession to European Economic Community. Land Use Policy 45, 18-25.
 https://doi.org/10.1016/j.landusepol.2015.01.004
- 630 Ćosović, M., Bugalho, M.N., Thom, D., Borges, J.G., 2020. Stand structural characteristics are
 631 the most practical biodiversity indicators for forest management planning in Europe.
 632 Forests 11, 343; <u>https://doi.org/10.3390/f11030343</u>
- Deus, E., Silva, J. S., Castro-Díez, P., Lomba, A., Ortiz, M. L., Vicente, J., 2018. Current and
 future conflicts between eucalypt plantations and high biodiversity areas in the Iberian
 Peninsula. Journal for Nature Conservation 45, 107-117,
 <u>https://doi.org/10.1016/j.jnc.2018.06.003</u>
- Díaz-Balteiro, L., Bertomeu, M., Bertomeu, M., 2009. Optimal harvest scheduling in
 Eucalyptus plantations: A case study in Galicia (Spain). Forest Policy and Economics
 11, 548-554. <u>https://doi.org/10.1016/j.forpol.2009.07.005</u>
- 640 Diéguez-Aranda, U., Rojo Alboreca, A., Castedo-Dorado, F., Álvarez González, J.G., Barrio641 Anta, M., Crecente-Campo, F., González González, J.M., Pérez-Cruzado, C., Rodríguez
 642 Soalleiro, R., López-Sánchez, C.A., Balboa-Murias, M.A., Gorgoso Varela, J.J.,
- 643 Sánchez Rodríguez, F., 2012. Herramientas selvícolas para la gestión forestal sostenible
 644 en Galicia. Adenda A. Actualización de modelos disponibles a fecha de 29/10/2012.
 645 Xunta de Galicia.
- 646 DOG, 2012. Ley 7/2012, de 28 de junio, de montes de Galicia. DOG nº 140, lunes 23 de julio
 647 de 2012. <u>https://xeg.xunta.gal/es/normativa/ley-72012-del-28-de-junio-de-montes-de-</u>
 648 <u>galicia</u>
- DOG, 2014. DECRETO 167/2019, de 5 de diciembre, por el que se crea y se regula el Registro
 de Masas Consolidadas de Frondosas Autóctonas. DOG nº 62, lunes 31 de marzo de

- 651 2014. <u>https://www.xunta.gal/dog/Publicados/2014/20140331/AnuncioCA02-270314-</u>
- 652 <u>0001_es.html</u>
- DOG, 2019. ORDE do 25 de novembro de 2019 pola que se establecen as bases reguladoras 653 654 das axudas para a creación de superficies forestais, cofinanciadas co Fondo Europeo 655 Agrícola de Desenvolvemento Rural (Feader) no marco do Programa de 656 desenvolvemento rural de Galicia 2014-2020, e se convocan para o ano 2020 (código 657 de procedemento MR670B). DOG nº 232, 5 de decembro de 2019. 658 https://www.xunta.gal/dog/Publicados/2019/20191205/AnuncioG0426-251119-
- 659 <u>0003_gl.html</u>

DOG, 2020. DECRETO 167/2019, de 5 de diciembre, por el que se crea y se regula el Registro
de Masas Consolidadas de Frondosas Autoctonas. DOG nº 4, miércoles 8 de enero de
2020. https://www.xunta.gal/dog/Publicados/2020/20200108/ AnuncioG0426-2712190009 es.html.

- Echeverría, C., Newton, A.C., Lara, A., Rey Benayas, J.M., Coomes, D., 2007. Impacts of forest
 fragmentation on species composition and forest structure in the temperate landscape of
 southern Chile. lobal Ecology and Biogeography 16, 426-439.
 https://doi.org/10.1111/j.1466-8238.2007.00311.x
- EEA, European Environmental Agency, 2019. CORINE Land Cover 2018. Available online:
 https://land.copernicus.eu/pan-european/corine-land-cover/clc2018 (accessed 15
 January 2020).
- FAO, 2010. Planted forests in sustainable forest management. A statement of principles.
 http://www.fao.org/3/al248e/al248e00.pdf
- 673 FAO, 2020. Global Forest Resources Assessment 2020. Main Report.
 674 http://www.fao.org/3/ca9825en/CA9825EN.pdf

- Freer-Smith, P., Muys, B., Bozzano, M., Drössler, L., Farrelly, N., Jactel, H., Korhonen, J.,
 Minotta, G., Nijnik, M. and Orazio, C., 2019. Plantation forests in Europe: challenges
 and opportunities. From Science to Policy 9. European Forest Institute.
 https://doi.org/10.36333/fs09
- García-Villabrille, J.D., Crecente-Campo, F., Pérez-Cruzado, C., Rodríguez-Soalleiro, R.,
 Diéguez-Aranda, U., Rojo-Alboreca, A., 2014. Biomass and carbon content in Galicia
 (NW Spain) *Eucalyptus globulus* Labill. stands. Recursos Rurais 10, 45-52
- García-Villabrille, J.D., 2015. Modelización del crecimiento y la producción ' de plantaciones
 de *Eucalyptus globulus* Labill. en el noroeste de España. Doctoral Thesis. Universidad
 de Santiago de Compostela. https://minerva.usc.es/xmlui/ handle/10347/13686.
- Gómez-García, E., 2020. Estimating the changes in tree carbon stocks in Galician forests (NW
 Spain) between 1972 and 2009. Forest Ecology and Management, 467, 118157.
 <u>https://doi.org/10.1016/j.foreco.2020.118157</u>
- Gonçalves, C.I., Vilas-Boas, L., Branco, M., Rezende, G.D., Valente, C., 2019. Host
 susceptibility to *Gonipterus platensis* (Coleoptera: Curculionidae) of *Eucalyptus*species. Ann. For. Sci. 76, 63. <u>https://doi.org/10.1007/s13595-019-0850-y</u>
- González-García, M., Hevia, A., Majada, J., Calvo de Anta, R., Barrio-Anta, M., 2015.
 Dynamic growth and yield model including environmental factors for *Eucalyptus nitens*(Deane & Maiden) Maiden short rotation woody crops in Northwest Spain. New Forests
 volume 46, 387-407. DOI 10.1007/s11056-015-9467-7.
- Guisan, A., Zimmermann, N.E., 2000. Predictive habitat distribution models in ecology. Ecol
 Model 135, 147-186. https://doi.org/10.1016/ S0304-3800(00)00354-9
- Hanski, I., 1999. Metapopulation Ecology. Oxford University Press, Oxford, UK.
- Hargis, C.D., Bissonette, J.A., David, J.L., 1998. The behavior of landscape metrics commonly
 used in the study of habitat fragmentation. Landscape Ecology 13, 167-186

- Harper, K.A., MacDonald, S.E., Burton, P., Chen, J., Brosofske, K.D., Saunders, S.,
 Euskirchen, E., Roberts, D., Jaiteh, M., Esseen, P., 2005. Edge influence on forest
 structure and composition in fragmented landscapes. Conservation Biology 19, 768703 782.
- Harrison, S., Bruna, E., 1999. Habitat fragmentation and large scale conservation:what do we
 know for sure? Ecography 22, 225-232. <u>https://doi.org/10.1111/j.1600-</u>
 0587.1999.tb00496.x
- Jactel, H., Koricheva, J., Castagneyrol, B., 2019. Responses of forest insect pests to climate
 change: not so simple. Current Opinion in Insect Science 35, 103–108.
 https://doi.org/10.1016/j.cois.2019.07.010
- 710 Kroner, R.E.G., Krithivasan, R., Mascia, M.B., 2016. Effects of protected area downsizing on
- habitat fragmentation in Yosemite National Park (USA), 1864–2014. Ecol. Soc. 21(3),
 22. <u>http://dx.doi.org/10.5751/ES-08679-210322</u>
- Laurance, W.F., 2000. Do edge effects occur over large spatial scales? Trends in Ecology and
 Evolution 15, 134-135
- Levenson, J.B., 1981. Woodlots as biogeographic islands in Southeastern Wisconsin. In
 Burgess RL and Sharpe DM (eds.). Forest island dynamics in man-dominated
 landscapes. New York: Springer-Verlag.
- Levers, C., Verkerk, P.J., Müller, D., Verburg, P.H., Butsic, V., Leitão, P.J., Lindner, M.,
 Kuemmerle, T., 2014. Drivers of forest harvesting intensity patterns in Europe. Forest
 Ecology and Management 315, 160-172. <u>https://doi.org/10.1016/j.foreco.2013.12.030</u>
- Lomba, A., Vicente, J., Moreira, F., Honrado, J., 2011. Effects of multiple factors on plant
 diversity of forest fragments in intensive farmland of Northern Portugal. Forest Ecology
 and Management, 262, 2219–2228. <u>https://doi.org/10.1016/j.foreco.2011.08.014</u>

- López-Sánchez, C.A., Castedo-Dorado, F., Cámara-Obregón, A., Barrio-Anta, M., 2021.
 Distribution of *Eucalyptus globulus* Labill. in northern Spain: contemporary cover,
 suitable habitat and potential expansion under climate change. Forest Ecology and
 Management 481, 118723. https://doi.org/10.1016/j.foreco.2020.118723
- Martín Gil, A., Mansilla Vázquez, J.P., Pérez Otero, R., 2018. Guía de gestión integrada de
 plagas de eucalipto. Ministerio de Agricultura, Pesca y Alimentación, 83 pp, Madrid.
- MAPA, 2019a. Spanish National Fourth Inventory updating. Ministerio de Agricultura, Pesca
 y Alimentación. Gobierno de España (Unpublished).
- MAPA, 2019b. Anuario de estadística Forestal. Ministerio de Agricultura, Pesca y
 Alimentacion. Gobierno de España. https://www.mapa.gob.es/es/desarrollo-rural/
 estadisticas/forestal anuarios todos.aspx.
- Marey Pérez, M.F., Fernández Alonso, S., Crecente Maseda, R., 2006. Planificación forestal
 táctica y ordenación territorial: planes de ordenación de los recursos forestales, una
 aplicación en el distrito ambiental I de Galicia (España). Recursos Rurais 2(1), 41-50
- Marey-Pérez, M.F., Rodríguez-Vicente, V., 2009. Forest transition in Northern Spain: Local
 responses on large-scale programs of field-afforestation. Land Use Policy 26(1),139-
- 740 156. https://doi.org/10.1016/j.landusepol.2008.02.004
- McGarigal, K., Marks, B.J., 1994. Spatial pattern analysis program for quantifying landscape
 structure. USDA Forest Service General Technical Paper. PNW–351.
- McGarigal, K., Wan, H.Y., Zeller, K.A., Timm, B.C., Cushman, S.A., 2016. Multi-scale habitat
 selection modeling: a review and outlook. Landsc. Ecol. 31, 1161-1175.
 <u>https://doi.org/10.1007/s10980-016-0374-x.</u>
- 746 McShea, W., 2014. What are the roles of species distribution models in conservation planning?.
- 747EnvironmentalConservation41(2),93-96.
- 748 https://doi.org/10.1017/S0376892913000581

749	Mendes, A.C., 2007. Uma historia de ascensao e queda. Arvores e florestas de Portugal, pinhais
750	e eucaliptais – a floresta cultivada (An history of rising and falling. Trees and forests of
751	Portugal, pine and eucalipt forests - the cultivated forest). Lisboa: Publico,
752	Comunicacao Social, SA and Fundacao Luso-Americana para o Desenvolvimento.
753	Midha, N., Mathur, P.K., 2010. Assessment of forest fragmentation in the conservation priority
754	Dudhwa landscape, India using FRAGSTATS computed class level metrics. J Indian
755	Soc Remote Sens 38, 487-500. https://doi.org/10.1007/s12524-010-0034-6
756	MITECO, 2019. Cartografía de Espacios naturales Protegidos de España. Ministerio para la
757	Transición Ecológica y el reto Demográfico.
758	https://www.miteco.gob.es/es/biodiversidad/temas/espacios-protegidos/espacios-
759	naturales-protegidos/default.aspx
760	MMA, 2002. Plan Forestal Español (Spanish Forest Plan).
761	https://www.mapa.gob.es/es/desarrollo-rural/temas/politica-forestal/planificacion-
762	forestal/politica-forestal-en-espana/pfe_descargas.aspx
763	Montero de Burgos, J.L., 1990. El eucalipto en España. Comentarios a un problema. ICONA,
764	Serie Técnica, 44 pp, Madrid.
765	Murcia, C., 1995. Edge effects in fragmented forests: implications for conservation. Trends in
766	Ecology & Evolution, 10(2), 58-62. <u>https://doi.org/10.1016/S0169-5347(00)88977-6</u>
767	Ónega-López, F.J., Puppim de Oliveira, J.A., Crecente-Maseda, R., 2008. Planning Innovations
768	in Land Management and Governance in Fragmented Rural Areas: Two Examples from
769	Galicia (Spain). European Planning Studies, 18(5): 755-773.
770	https://doi.org/10.1080/09654311003594067
771	Palik, B.J., Murphy, P.G., 1990. Disturbance verses edge effects in sugar-maple/beech forest

- Parrotta, J.A., Turnbull, J.W., Jones, N., 1997. Catalyzing native forest regeneration on
 degraded tropical lands, Forest Ecology and Management 99(1-2), 1-7.
 https://doi.org/10.1016/S0378-1127(97)00190-4
- Pawson, S.M., Brin, A., Brockerhoff, E.G., Lamb, D., Payn, T.W., Paquette, A., Parrotta, J.A.,
 2013. Plantation forests, climate change and biodiversity Biodivers. Conserv. 22(5),
 1203-1227
- Pérez, S., Renedo, C.J., Ortiz, A., Mañana, M., Silió, D., 2006. Energy evaluation of the *Eucalyptus globulus* and the *Eucalyptus nitens* in the north of Spain (Cantabria).
 Thermochimica Acta 451(1-2), 57e64. <u>https://doi.org/10.1016/j.tca.2006.08.009</u>
- Pérez-Cruzado, C. 2011. Models for estimating biomass and carbon in biomass and soils in
 Pinus radiata (D. Don), *Eucalyptus globulus* (Labill) and *Eucalyptus nitens* (Deane &
- Maiden) Maiden plantations established in former agricultural lands in northwestern
 Spain. Doctoral Thesis. Universidad de Santiago de Compostela.
- 786 https://minerva.usc.es/xmlui/handle/10347/3674
- Prado, J.A., Barros, S., 1989. *Eucalyptus*, Principios de silvicultura y manejo. Instituto ForestalCORFO. Santiago de Chile, 199 p.
- Ranney, J.W., Bruner, M.C., Levenson., 1981. The important of edge in the structure and
 dynamics of forest islands. In Burgess RL and Sharpe DM (eds.). Forest island dynamics
 in man-dominated landscapes. New York: Springer-Verlag.
- Rico, J., 2018. Galicia quiere jugar la liga de campeones de la biomasa. Energías renovables,
 176, 52-54.
- Riesco, G., 2004. Forest management in *Eucalyptus* stands: The Spanish case. In International
 IUFRO Meeting of the WP4.04.06 on planning and economics of fast-growing
 plantation forests. The economics and management of high productivity plantations,
 Lugo, Galicia (Spain), 27–30 September 2004. Edited by Álvarez Gozález, J.G.;

- Goulding, C., Rojo, A., Rodríguez, R., Zoralioglu, T., Ruiz, A.D. University of Santiago
 de Compostela, Spain.
- Robak, E., Aboal, J., Picos, J., 2012. Sustainable Forest Management in Galicia (Spain):
 Lessons Learned. In: Sustainable Forest Management Case Studies. Edited by: MartinGarcia, J., Diez, J.J. IntechOpen, 221-238. <u>https://doi.org/10.5772/29706</u>
- SCBD, 2005. Handbook of the Convention on Biological Diversity Including its Cartagena
 Protocol on Biosafety. Secretariat of the Convention on Biological Diversity, 3rd
 edition, (Montreal, Canada).
- Sinclair, S.J., White, M.D., Newell, G.R., 2010. How Useful Are Species Distribution Models
 for Managing Biodiversity under Future Climates?. Ecol. Soc. 15(1), 8.
 https://doi.org/10.5751/ES-03089-150108
- Teixido, A.L., Quintanilla, L.G., Carreño, F., Gutiérrez, D., 2010. Impacts of changes in land
 use and fragmentation patterns on Atlantic coastal forests in northern Spain. Journal of
 Environmental Management 91, 879-886. DOI: 10.1016/j.jenvman.2009.11.004
- 812 Vaz, A.S., Honrado, J.P., Lomba, A., 2019. Replacement of pine by eucalypt plantations:
- 813 Effects on the diversity and structure of tree assemblages under land abandonment and
- 814 implications for landscape management. Landscape and Urban Planning 185, 61-67.
- 815 <u>https://doi.org/10.1016/j.landurbplan.2019.01.009</u>
- 816 Veiras, X., Soto, M.A., 2011. La conflictividad de las plantaciones de eucalipto en España y
 817 Portugal. 96 pp. Greenpeace, Madrid.
- Viera, M., Ruíz Fernandez, F., Rodríguez-Soalleiro, R., 2016. Nutritional prescriptions for
 eucalyptus plantations: lessons learned from Spain. Forests 7, 84.
 https://doi.org/10.3390/f7040084

821	Villero, D., Pla, M., Camps, D., Ruiz-Olmo, J., Brotons, L., 2017. Integrating species
822	distribution modelling into decision making to inform conservation actions. Biodivers
823	Conserv. 26, 251-271. https://doi.org/10.1007/s10531-016-1243-2

- Xu, J., Hou, H., Liu, B., Hu, J., 2017. The integration of different pretreatments and ionic liquid
 processing of eucalyptus: Hemicellulosic products and regenerated cellulose fibers.
 Industrial Crops and Products 101, 11-20.
 https://doi.org/10.1016/j.indcrop.2017.02.038
- Xunta de Galicia, 2018a. 1ª revisión del Plan Forestal de Galicia. Consellería do Medio Rural,
 Xunta de Galicia,75 pp.

Xunta de Galicia, 2018b. 1ª revisión del Plan Forestal de Galicia. Documento de diagnóstico
del monte y del sector forestal gallego. Consellería do Medio Rural, Xunta de
Galicia,195 pp.

Xunta de Galicia, 2020. Estadísticas Forestales. Consellería do Medio Rural, Xunta de Galicia.
 <u>https://mediorural.xunta.gal/es/recursos/estadisticas/estadisticas-forestales</u>

835 Yañez-S, M., Matsuhiro, B., Maldonado, S., González, R., Luengo, J., Uyarte, O., Serafine, D.,

- 836 Moya, S., Romero, J., Torres, R., Kogan, M.J., 2018. Carboxymethylcellulose from
- 837 bleached organosolv fibers of *Eucalyptus nitens*: synthesis and physicochemical
- 838 characterization. Cellulose 25, 2901-2914. <u>https://doi.org/10.1007/s10570-018-1766-7</u>

8. Tables

Feature		Metric	Symbol	Units	Description
Suitability ¹		Degree of suitability	DS	%	Ratio between the area of suitable E. globulus habitat included in the ForDis and the whole forest area in ForDis
Importance ²		Degree of occupancy	DO	%	Ratio between area covered by eucalypt in the ForDis and the total wooded area
Importance		Economic relevance	ER	%	Ratio between mean annual volume of eucalypt harvested and the mean total volume harvested in the ForDis
		Potential for replacement	PR	%	Ratio between the current area of native broadleaved forest included in the <i>E. globulus</i> suitable habitat in the ForDis (SH_{NBF}) and the total area in the whole ForDis (S_{NBF})
Potential ¹ expansion		Potential for substitution	PS	%	Ratio between the current area of coniferous or mixed forest included in <i>E. globulus</i> suitable habitat in the ForDis (SH _{CMF}) and the total area in the whole ForDis (S _{CMF})
·		Potential for afforestation	PF	%	Ratio between the area of natural and semi-natural non-wooded areas included in the <i>E. globulus</i> suitable habitat in the ForDis (SH_{FNA}) and the total area in the whole ForDis (S_{FNA})
	a	Largest patch index	LPI	%	Proportion of the landscape encompassed by the largest patch
	Area	Mean Patch Size	MPS	ha	Average size of the patches of a forest type category
	₹,	Mean Shape Index	MSI	-	Average shape index of patches of corresponding forest type category, adjusted by a constant for a square standard
	ore	Total core area	TCA	ha	Sum of core areas of all patches corresponding to a forest type category
Fragmentation ¹	C	Mean core area	MCA	ha	Average core areas of the patches of the corresponding forest type category
1 ragmentation	ggregation	Patch density	PD	patches/1 00 ha	Number of patches in a category divided by total area and multiplied by 100
	Aggreg	Euclidean nearest neighbour distance	ENN	m	Average distance between patches of corresponding forest type category, based on the edge-to-edge distance

Table 1. Metrics used to characterize eucalypt plantations and native broadleaved forest.

841 ¹ Feature analysed for *E. globulus* suitable habitat. ² Feature analysed for whole ForDis.

844	Table 2 Area (ha)	of forest cover categories	at different levels of analysis.
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						Area in	the whole l	ForDis		-						Area i	ncluded in	suitable hab				
					Total area	ı				O	ily areas i NP	ncluded i As	n		Total	area		Only areas included in NPAs				
Province	ForDis	Sf	Sw	EUGL _{SNF4.5}	EUNI _{SNF4.5}	EU _{SNF4.5}	SNBF	SCMF	SFSA	EU _{SNF4.5}	SNBF	SCMF	SFSA	EU _{SNF4.5}	SHNBF	SHCMF	SHFSA	EU _{SNF4.5}	SHNBF	SHCMF	SHFSA	
	Ι	112252	85743	54419	1576	55995	23056	6692	26510	2745	4745	1070	6061	53866	21035	5886	19909	2633	4197	901	4936	
ŭa	П	113163	92919	43761	6533	50295	17442	25182	20245	148	365	267	605	49604	17008	23956	15599	146	365	258	563	
ruñ	III	127766	104603	33046	18405	51451	28543	24609	23164	383	436	586	3191	50788	27969	24055	19432	333	429	467	1588	
C	IV	66517	47008	16474	5100	21573	9747	15688	19509	59	149	169	276	21244	9593	14716	18580	58	149	164	268	
A	V	81921	49348	22209	2185	24394	7513	17441	32573	117	94	522	3244	23917	7340	15873	29493	115	90	496	3210	
	Subtotal	501620	379620	169908	33800	203708	86301	89611	122001	3452	5790	2614	13379	199419	82945	84485	103013	3285	5231	2287	10565	
	VI	110492	91505	60689	3018	63707	19253	8545	18988	1598	1839	1702	5660	62045	17421	6415	13440	1468	1196	729	2908	
	VII	142269	83605	251	1127	1378	64190	18037	58663	75	26682	4397	26280	0	0	0	0	0	0	0	0	
ogu	VIII	133205	73990	94	71	165	42606	31219	59214	0	11295	4540	20277	1	68	12	68	0	25	12	39	
Γ	IX	137443	99639	941	6123	7064	59030	33545	37805	8	5330	430	5157	584	1074	986	1849	4	42	84	359	
	Х	123801	77978	3861	12224	16085	35345	26548	45823	328	3356	1025	10385	4651	10199	4619	11912	53	640	199	1999	
	Subtotal	647209	426714	65835	22563	88398	220423	117893	220494	2010	48503	12094	67759	67281	28763	12032	27268	1524	1903	1023	5305	
	XI	78712	44907	359	55	414	22405	22088	33805	0	545	408	4044	197	7140	5264	6404	0	80	17	396	
se	XII	108657	57430	43	0	43	36552	20835	51227	0	1160	29	4698	3	514	655	806	0	0	0	28	
ren	XIII	133651	39124	0	6	6	32279	6839	94527	0	9487	1345	31325	0	0	0	0	0	0	0	0	
no	XIV	136074	38393	50	10	60	18699	19634	97681	0	3605	2418	25075	0	4	1	34	0	0	0	0	
-	XV Subtotal	89929 547023	26127 205981	128 580	30 100	157 680	19242 129177	6728 76124	63802 341042	8	2027 16824	875 5075	20162 85304	0 200	20 7679	4 5930	424 7667	0	16 96	17	384 807	
	XVI	94435	58363	5368	472	5839	30820	21704	36071	8 50	2785	983	85304	5521	19893	18080	16237	37	96 1619	699	2440	
ra	XVI	51407	31373	6213	956	7169	14976	9228	20034	25	2785	26	22	6660	13646	7715	14099	25	215	26	2440	
ved	XVII	58024	41289	11889	38	11927	11858	17504	16735	11	522	853	92	11576	11532	17123	14099	9	519	849	92	
ntev	XIX	86390	62282	27376	228	27605	21581	13096	24108	84	838	187	2800	26698	20962	1/123	20362	82	818	179	2713	
Por										170	4360	2049	11423	50455				152	3170	1753	5267	
	Subtotal	290255	193307	50847	1693	52540	79235	61532	96948						66034	55183	65635					
Note: Sf:	Total	1986106	1205622	287170	58156	345326	515136	345160	780484	5640	75477	21832	177864	317356	185420	157630	203584	4962	10401	5080	21944	

Table 3.- Current values of indicators of suitability, importance and potential expansion at ForDis and NPAs spatial levels.
849

			Tot	tal area					Areas incl	uded in Nl	PAs	
	Feature	Suitability	Abundance	Economic relevance	Suitability	Abundance	Potential expansion					
	Indicator	DS (%)	DO (%)	ER (%)	PR (%)	PS (%)	PF (%)	DS (%)	DO (%)	PR (%)	PS (%)	PF (%)
	ForDis											
	Ι	89.7	65.3	87.8	91.2	88.0	75.1	86.6	32.1	88.4	84.2	81.4
ña	II	93.8	54.1	74.9	97.5	95.1	77.1	96.2	18.9	100.0	96.7	93.0
Coruña	III	95.7	49.2	68.2	98.0	97.7	83.9	61.3	27.3	98.5	79.7	49.8
J	IV	96.4	45.9	62.6	98.4	93.8	95.2	97.9	15.6	100.0	96.9	97.2
¥	V	93.5	49.4	60.7	97.7	91.0	90.5	98.3	15.9	95.3	95.1	98.9
	Subtotal	93.7	53.7	75.0	96.1	94.3	84.4	84.7	29.1	90.3	87.5	79.0
	VI	89.9	69.6	83.8	90.5	75.1	70.8	58.3	31.1	65.1	42.8	51.4
	VII	0.0	1.6	19.3	0.0	0.0	0.0	0.0	0.2	0.0	0.0	0.0
Lugo	VIII	0.1	0.2	5.4	0.2	0.0	0.1	0.2	0.0	0.2	0.3	0.2
Lu	XIX	3.3	7.1	22.1	1.8	2.9	4.9	4.5	0.1	0.8	19.5	7.0
	Х	25.3	20.6	44.1	28.9	17.4	26.0	19.1	7.0	19.1	19.4	19.2
	Subtotal	20.9	20.7	47.4	13.0	10.2	12.4	7.5	3.2	3.9	8.5	7.8
	XI	24.1	0.9	18.1	31.9	23.8	18.9	9.9	0.0	14.7	4.1	9.8
e	XII	1.8	0.1	4.9	1.4	3.1	1.6	0.5	0.0	0.0	0.0	0.6
ens	XIII	0.0	0.0	2.5	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Ourense	XIV	0.0	0.2	3.7	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
0	XV	0.5	0.6	9.5	0.1	0.1	0.7	1.7	0.3	0.8	0.1	1.9
	Subtotal	3.9	0.3	7.9	5.9	7.8	2.2	0.9	0.0	0.6	0.3	0.9
в	XVI	63.3	10.0	31.8	64.5	83.3	45.0	38.9	1.3	58.1	71.0	28.7
dr	XVII	81.9	22.9	39.3	91.1	83.6	70.4	99.9	9.4	99.9	100.0	100.0
eve	XVIII	95.1	28.9	45.3	97.3	97.8	89.3	99.4	0.8	99.3	99.6	100.0
Pontevedra	XIX	92.9	44.3	70.4	97.1	93.7	84.5	97.0	7.6	97.6	95.8	96.9
Pc	Subtotal	81.8	27.2	49.0	83.3	89.7	67.7	57.5	2.6	72.7	85.6	46.1
	Total	43.5	28.6	58.4	36.0	45.7	26.1	15.1	5.5	13.8	23.3	12.3

	Category	Area													Core									Aggregation							
	Metric	LPI MPS MSI												ТСА МСА										PD	1	ENN					
	Forest type	(1)	(2)	(1) NPA	(2) NPA	(1)	(2)	(1) NPA	(2) NPA	(1)	(2)	(1) NPA	(2) NPA	(1)	(2)	(1) NPA	(2) NPA	(1)	(2)	(1) NPA	(2) NPA	(1)	(2)	(1) NPA	(2) NPA	(1)	(2)	(1) NPA	(2) NPA		
	Ι	25.7	1.6	1.2	7.5	44.6	3.0	6.3	9.0	1.8	1.3	1.5	1.5	43452.1	13159.3	1742.3	3145.3	35.8	1.9	4.2	6.7	0.9	5.0	2.9	3.2	116.7	86.4	168.6	116.5		
_	П	5.8	0.3	0.7	2.8	23.2	3.0	0.7	3.7	1.8	1.4	1.3	1.8	36925.3	9839.3	41.4	171.6	17.2	1.7	0.2	1.7	1.2	3.2	6.8	3.4	120.0	81.8	265.1	84.9		
Coruña	III	1.0	0.4	2.5	0.7	16.1	5.4	1.7	1.5	1.8	1.5	1.3	1.7	35703.8	18430.3	190.3	137.3	11.2	3.6	1.0	0.5	1.4	2.3	4.4	6.8	122.1	80.8	289.5	134.6		
A C	IV	1.7	1.1	1.5	6.4	21.1	4.2	1.3	4.6	1.9	1.4	1.2	1.7	15418.4	6260.6	29.6	87.5	15.2	2.7	0.7	2.7	1.1	2.6	3.1	2.4	118.6	81.4	143.7	128.4		
	v	2.1	0.4	0.2	0.8	16.9	4.4	2.5	3.8	1.9	1.4	1.5	1.5	16799.1	4768.3	49.7	47.8	11.8	2.9	1.1	2.0	1.2	1.4	1.1	0.5	123.9	93.7	658.3	789.8		
	Subtotal	7.2	0.8	1.2	3.6	24.3	4.0	2.5	4.5	1.9	1.4	1.4	1.6	29659.7	10491.5	410.7	717.9	18.3	2.6	1.4	2.7	1.2	2.9	3.6	3.3	120.3	84.8	305.1	250.8		
	VI	39.8	1.1	3.6	2.9	74.6	3.0	3.6	3.0	1.8	1.3	1.3	1.5	51409.0	10761.0	1043.7	690.3	61.7	1.9	2.6	1.7	0.7	4.7	5.6	5.6	102.2	91.6	141.0	134.4		
	VII*	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-		
Lugo	VIII	0.1	1.6	0.0	26.8	0.6	1.9	0.0	3.6	1.5	1.3	0.0	1.3	0.0	29.9	0.0	15.1	0.0	0.8	0.0	2.2	0.2	2.8	0.0	8.8	26226.7	657.5	0.0	107.0		
Г	IX	0.7	1.5	0.4	4.0	3.0	4.8	0.7	2.7	1.5	1.5	1.3	1.4	287.9	637.8	0.9	23.4	1.5	2.8	0.2	1.6	2.6	3.0	1.0	2.9	426.3	361.3	1775.1	331.5		
	Х	0.4	0.6	0.2	1.9	4.3	7.4	1.2	4.0	1.5	1.5	1.3	1.6	2563.0	6689.9	18.6	352.4	2.3	4.9	0.4	2.2	2.0	2.5	1.2	4.6	256.4	146.1	1281.6	354.9		
	Subtotal	10.3	1.2	1.4	8.9	20.6	4.3	1.8	3.3	1.6	1.4	1.3	1.5	13565.0	4529.7	354.4	270.3	16.4	2.6	1.1	1.9	1.4	3.3	2.6	5.5	6752.9	314.1	1065.9	231.9		
	XI	0.1	2.7	0.0	7.1	3.5	17.6	0.0	4.5	1.5	1.6	0.0	1.5	104.0	5274.1	0.0	50.9	1.9	12.9	0.0	2.8	0.2	1.7	0.0	3.6	907.6	205.1	0.0	1417.1		
	XII	0.1	3.4	0.0	0.5	2.8	10.1	0.0	0.1	1.1	1.6	0.0	1.0	1.5	348.0	0.0	0.0	1.5	6.8	0.0	0.0	0.0	1.6	0.0	3.7	-	483.0	0.0	0.0		
Ourense	XIII	0.0	3.1	0.0	0.0	0.0	1.5	0.0	0.0	0.0	1.1	0.0	0.0	0.0	1.9	0.0	0.0	0.0	0.7	0.0	0.0	0.0	3.4	0.0	0.0	0.0	3262.1	0.0	0.0		
Ō	XIV*	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-	-		
	XV	0.0	0.7	0.0	0.6	0.0	2.0	0.0	2.3		1.3	0.0	1.3	0.0	8.8	0.0	6.9	0.0	0.9	0.0	1.0	0.0		0.0	1.1	0.0	2791.2	0.0	3890.7		
	Subtotal	0.1	2.5	0.0	2.7	3.1	7.8	0.0	2.3		1.4	0.0	1.2	52.8	1408.2	0.0	19.3	1.7	5.3	0.0	1.3	0.1		0.0	2.8	907.6	1685.3	0.0	2653.9		
	XVI	1.1	2.9	0.4	9.0	9.7	20.5	2.6	10.4		1.6	1.3	1.7	3594.3	15401.3	19.8	1174.6	6.3	15.8	1.4	7.5			0.3	2.8	281.5	143.9	2787.0	173.2		
edra	XVII	0.8	5.5	1.7	15.7		14.3	0.7	1.6		1.5	1.3	1.5	4798.9	10220.7	6.6	61.2	13.2		0.2	0.4		1.7	10.6	37.6	157.9	81.7	359.4	98.9		
Pontevedra	XVIII	2.2	1.3	0.1	9.8	20.0		0.5	32.7		1.4	1.2	2.0	8433.3	8501.2	2.1	422.1	14.5	5.8	0.1	26.4	0.7		0.8	0.8	189.8	82.9	755.1	1924.4		
Ā	XIX	2.0	1.3	0.4	5.4	25.0	6.0	1.6	16.8		1.5	1.4	2.0	19936.6	13942.7	32.9	553.5	18.6	4.0	0.6	11.3	0.9		1.2	1.1	139.8	77.2	771.5	1016.8		
	Subtotal	1.5	2.8	0.7	10.0	18.3		1.3	15.4		1.5	1.3	1.8	9190.8	12016.5	15.3	552.9	13.1	9.1	0.6	11.4			3.2	10.6	192.3	96.4	1168.2	803.3		
	TOTAL	5.6	1.7	1.1	6.4	18.9	6.9	2.0	6.5	1.7	1.4	0.0	1.6	15961.8	7310.3	264.8	433.8	14.2	4.7	1.1	4.4	1.0	2.5	3.2	5.6	2092.1	518.1	783.0	713.5		

Table 4. Fragmentation metrics for eucalypt plantations and native broadleaved forest included in the current *E. globulus*-suitable habitat.

(1) *E. globulus* + *E. nitens* (EU_{SNF4.5}); (2) Native broadleaved forest (SH_{NBF}); Natural Protected Areas (NPA). * ForDis without surface of *E. globulus* suitable habitat.

Table 5- Forecasted future gains or losses (percentage) of suitability and potential expansion indicators under two future climate change scenarios.

					Total	area				Areas included in NPAs												
		2050 un climate			-	2050 un climate					50 under 1 nate chang)	2050 under pessimistic climate change scenario								
	Feature	Suitability	e	Potentia xpansio	n	Suitability	e	Potentia <u>xpansio</u>	n	Suitability		Potential <u>expansion</u>		Suitability	Potential expansion							
	Indicator	DS (%)	PR (%)	PS (%)	PF (%)	DS (%)				DS (%)	PR (%)	PS (%)	PF (%)	DS (%)	PR (%)	PS (%)	PF (%)					
	Ι	4.7	5.8	9.9	12.4	2.9	5.8	9.9	12.4	11.7	11.4	15.0	16.7	11.7	11.4	15.0	16.7					
ía	II	2.1	1.9	1.9	7.7	0.1	1.8	2.0	8.1	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0					
Coruña	III	2.6	1.8	1.4	10.8	4.1	2.0	2.1	15.4	23.2	1.5	12.5	30.9	34.6	1.5	16.1	46.7					
	IV	0.0	0.0	0.0	0.0	-1.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0					
A	V	0.1	0.2	0.2	0.0	-1.8	0.2	0.2	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0					
	Subtotal	2.2	2.5	1.7	6.0	1.3	2.6	1.9	7.0	11.0	9.5	8.9	14.9	13.1	9.5	9.7	18.7					
	VI	2.9	1.1	12.0	10.4	-7.6	0.0	6.9	3.4	15.6	6.6	37.3	16.4	5.4	2.9	22.0	2.8					
	VII	0.5	0.6	0.7	0.2	0.3	0.3	0.4	0.3	0.6	0.8	1.2	0.2	0.2	0.2	0.5	0.1					
Lugo	VIII	5.5	5.9	6.6	4.6	22.6	24.2	28.7	18.3	0.9	0.7	2.3	0.8	6.2	4.8	16.9	4.6					
Lu	IX	33.6	32.8	43.0	32.9	39.0	35.0	50.4	36.2	9.5	13.2	19.8	4.8	10.5	13.8	20.0	6.3					
	Х	45.8	57.6	62.0	43.2	47.4	52.2	61.0	36.9	34.7	69.3	31.6	24.9	19.2	59.0	22.8	6.7					
	Subtotal	17.6	19.4	28.9	16.8	15.4	22.5	36.3	19.2	6.6	7.1	9.9	5.9	5.3	7.0	12.3	3.1					
	XI	57.5	58.3	62.6	54.4	70.2	65.3	75.0	70.7	40.6	35.9	50.9	40.2	53.5	46.7	56.4	54.2					
e	XII	30.3	26.7	49.5	25.1	60.2	58.4	86.6	50.7	7.2	9.6	11.5	6.6	62.3	73.2	49.3	59.7					
ens	XIII	0.1	0.5	0.0	0.0	0.7	1.8	1.2	0.3	0.1	0.4	0.0	0.0	0.4	1.3	0.0	0.2					
Ourense	XIV	0.0	0.0	0.0	0.0	3.6	4.0	3.1	3.6	0.0	0.0	0.0	0.0	0.6	4.4	0.0	0.1					
Ŭ	XV	5.0	7.9	12.7	3.3	22.9	26.4	25.7	21.6	3.6	11.9	6.2	2.7	38.8	52.5	37.3	37.5					
	Subtotal	15.1	19.0	32.8	9.8	26.8	32.8	48.6	19.8	3.1	3.5	5.2	2.9	14.6	14.6	11.3	14.8					
æ	XVI	26.6	29.4	13.7	36.3	30.9	32.4	14.7	44.4	41.9	35.1	15.4	47.4	53.6	38.0	17.6	63.1					
adr	XVII	13.0	5.1	10.6	24.6	15.5	5.1	10.6	29.5	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0					
teve	XVIII	2.2	0.0	0.0	7.6	2.3	0.0	0.0	7.8	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0					
Pontevedra	XIX	3.0	0.4	0.0	10.2	2.6	0.4	0.0	10.4	2.6	2.4	0.0	3.0	2.6	2.4	0.0	3.0					
Р	Subtotal	12.3	12.5	6.4	22.4	14.0	13.7	6.8	26.5	29.2	22.9	7.4	36.0	37.3	24.7	8.5	47.8					
	Total	12.3	15.4	18.7	12.7	10.2	20.4	24.8	18.4	11.7	11.4	15.0	16.7	11.7	11.4	15.0	16.7					

859 9. Figure Captions

860 Figure 1. Location of the study area (Galicia) within the Iberian Peninsula, showing the provinces and

- the forest planning units (forest districts are indicated by Roman numerals). Forest districts: I.-Ferrol;
- 862 II.-Bergantiños-Mariñas Coruñesas; III.- Santiago-Meseta Interior; IV.- O Barbanza; V.- Fisterra; VI.-
- 863 A Mariña Luguesa; VII.- A Fonsagrada-Os Ancares; VIII.- Terra de Lemos; IX.- Lugo-Sarria; X.- Terra
- 864 Chá; XI.- O Riberiro-Arenteiro; XII.-Miño-Arnoia; XIII.- Valdeorras-Trives; XIV.- Verín-Viana; XV.-
- A Limia; XVI.- Deza-Tabeirós; XVII.- O Condado-A Paradanta; XVIII.- Vigo-Baixo Miño; XIX.Caldas-O Salnés.
- 867
- 868 Figure 2. Workflow of the different processes used in the study.
- 869

Figure 3. Spatial distribution at ForDis level of eight metrics used to characterize eucalypt plantations
and native broadleaved forest. DS=degree of suitability; DO=degree of occupancy; ER=economic
relevance; PR=potential for replacement; PS=potential for substitution; PF=Potential for afforestation;
LPI_EU=large patch index of eucalypt plantations; LPI_NBF=large patch index of native broadleaved
forests; MPS_EU=mean patch size of eucalypt plantations; MPS_NBF=mean patch size of native
broadleaved forest.

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Figure 4. Map showing the large area of eucalypt plantations in ForDis VI- A Mariña Lucense (upper
right) and the expansion of *E. nitens* in the ForDis X - Terra Chá (lower right).

879 S_{FSA}= area occupied by other natural and semi-natural land. S_{CMF}= area occupied by coniferous and

- 880 mixed forest. S_{NBF} =area occupied by native broadleaved forest. EUGL_{SNF4.5}= area occupied by *E*.
- 881 *globulus* plantations. EUNI_{SNF4.5}= area occupied by *E. nitens* plantations.



















